This file has been cleaned of potential threats.

If you confirm that the file is coming from a trusted source, you can send the following SHA-256 hash value to your admin for the original file.

78838bc21307c5028073cde93dec74176abedb20411f68937a3bb7dac8c3e5b4

To view the reconstructed contents, please SCROLL DOWN to next page.

# Environmental Conservation



cambridge.org/enc

## **Research Paper**

**Cite this article:** Koga DM et al. (2022). Serra do Divisor National Park: a protected area under threat in the south-western Brazilian Amazon. *Environmental Conservation* **49**: 74–82. doi: 10.1017/S0376892922000091

Received: 27 July 2021 Revised: 16 January 2022 Accepted: 25 February 2022

Keywords: Amazonia; Brazil; deforestation; land use; remote sensing; roads

Author for Correspondence: Diogo Mitsuru Koga, Email: diogomkoga@hotmail.com

© The Author(s), 2022. Published by Cambridge University Press on behalf of Foundation for Environmental Conservation. This is an Open Access article, distributed under the terms of the Creative Commons Attribution licence (https:// creativecommons.org/licenses/by/4.0/), which permits unrestricted re-use, distribution, and reproduction in any medium, provided the original work is properly cited.



# Serra do Divisor National Park: a protected area under threat in the south-western Brazilian Amazon

Diogo Mitsuru Koga<sup>1</sup>, Irving Foster Brown<sup>2,3</sup>, Philip M Fearnside<sup>1</sup>, David S Salisbury<sup>4</sup> and Sonaira Souza da Silva<sup>3</sup>

<sup>1</sup>Instituto Nacional de Pesquisas da Amazônia, Programa de Pós-Graduação em Gestão de Áreas Protegidas na Amazônia, Manaus, AM, CEP 69067-375, Brazil; <sup>2</sup>Woodwell Climate Research Center, 149 Woods Hole Rd, Falmouth, MA 02540, USA; <sup>3</sup>Universidade Federal do Acre, Geoprocessing Laboratory Applied to the Environment – LabGAMA, Cruzeiro do Sul, AC, CEP 69980-000, Brazil and <sup>4</sup>University of Richmond, Department of Geography and the Environment, Westhampton Way, Richmond, VA 23173, USA

## Summary

Protected areas have numerous roles (such as biodiversity preservation, the development of scientific research and the sustainable use of natural resources), but they are under threat from political and economic forces. The 837 000-ha Serra do Divisor National Park (SDNP) in the south-western Brazilian Amazon combines the conservation of natural resources and the maintenance of the productive activities of *c*. 400 resident families. The Brazilian and Peruvian governments have proposed a road linking Acre (Brazil) to Ucayali (Peru) that would bisect the SDNP. Another threat to the SDNP is a bill proposing its downgrading to an 'environmental protection area'. This study aims to map the land cover of the SDNP and its surroundings from 1988 to 2018 and to analyse the dynamics of land-use change. Analysis of Landsat satellite images with supervised classification using the MaxVer algorithm show that, during the 30-year period, pasture showed the highest absolute land-cover gain, with 1986 ha in the interior and 7661 ha along the periphery of the SDNP. Only 1% of the park's primary forest was lost by 2018, but the proposed road and potential downgrading may result in accelerated deforestation and forest degradation in the near future.

#### Introduction

In an era of increasing human pressure on ecosystems and biodiversity, protected areas (PAs) have emerged as a pillar of conservation efforts (Nelson & Chomitz 2011, Barber et al. 2014). PAs, which include conservation units (CUs), Indigenous lands (ILs) and military areas, are home to 54% of the remaining forests in the Brazilian Amazon and contain 56% of their forest carbon (Soares-Filho et al. 2010).

Jenkins and Joppa (2009) analysed the worldwide creation of new PAs after 1985 and found that between 2003 and 2009 Brazil created 74% (523 592 km<sup>2</sup>) of the total area (703 864 km<sup>2</sup>) of new PAs on the planet. The locations of PAs are fixed, but impacts can reach the most isolated locations due to environmental changes from regional and/or global degradation and climate change (Wiens et al. 2011). PAs near roads are at greatest risk for deforestation in the Amazon (Barber et al. 2014), where the Serra do Divisor National Park (SDNP) has come under threat due to the proposed Cruzeiro do Sul–Pucallpa road (Ruaro & Laurance 2022). The SDNP created by Federal Decree No. 97,839 of 16 June 1989 is one of the world's most biodiverse national parks and contains many endemic species (Whitney et al. 2004, Silveira et al. 2008, Dolibaina et al. 2015, Bernarde et al. 2016).

Infrastructure expansion projects in the Amazon present one of the main threats to PAs, and their presence can accelerate the political and economic forces that relax the level of protection of these areas in the Brazilian Amazon (Ferrante & Fearnside 2020, Fearnside 2021). Many conflicts result from the proximity of PAs to major infrastructure projects, such as the BR-163 and BR-319 highways and the Belo Monte hydroelectric dam (Fleury & Almeida 2013, Barni et al. 2015). The Brazilian legislative bill 6024/2019 currently threatens the SDNP with a reclassification from a national park to an 'area of environmental protection' (APA – Área de Proteção Ambiental). Reclassification as an APA would allow natural resource extraction currently not permitted in a national park. Developers are particularly interested in the extremely rare rock in Acre state from the Serra do Divisor mountains that would facilitate the construction of the proposed highway between Cruzeiro do Sul (Brazil) and Pucallpa (Peru) and the maintenance of this and other roads throughout Acre (Fig. 1). On the Peruvian side, the proposed road would cross the SDNP's sister park, the Sierra del Divisor

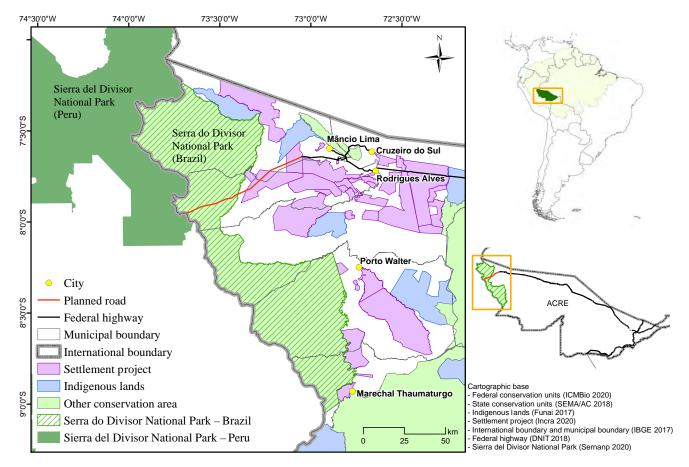


Fig. 1. Location and limits of Brazil's Serra do Divisor National Park.

National Park, which was created by Peruvian Supreme Decree n° 014-2015-MINAM to protect the park area's high levels of biodiversity and unique topographical features.

The Cruzeiro do Sul-Pucallpa transboundary road project, if carried out, would be the second highway from Brazil that goes through Peru to reach the Pacific Ocean, and it would be a key link in the Central Interoceanic Route (abbreviated as 'IOC' in Spanish), which would be an alternative to the recently paved Southern Interoceanic Route (abbreviated as 'IOS' in Spanish; Hopkins et al. 2015). Both roads (the existing southern route and the proposed central route) are part of the Initiative for the Integration of Regional Infrastructure in South America (IIRSA), which is best known by this name despite having been officially renamed the South American Council for Infrastructure and Planning (COSIPLAN). The IIRSA's goal is to improve transportation, energy and telecommunications infrastructure corridors across South America to stimulate economic growth and to reduce poverty. These initiatives have accessed remote rural areas, contacted diverse Amazonian cultures and threatened intact ecosystems (Killeen 2007).

The present study examines the dynamics of land use and land cover in the SDNP and its surrounding buffer zone over the 30 years from 1988 to 2018. The goal is to analyse the effectiveness of the SDNP in maintaining forest and to anticipate the implications and consequences of the proposed road and of the potential downgrading of the national park to an environmental protection area.

### Methods

#### Study area

The SDNP (Fig. 1) comprises 837 599 ha in the municipalities (counties) of Cruzeiro do Sul, Mâncio Lima, Rodrigues Alves, Marechal Thaumaturgo and Porto Walter. This area represents 5.5% of Acre state and runs along the Peruvian border in the south-western Brazilian Amazon.

#### Mapping land-use change between 1988 and 2018

The mapping of land use was carried out using Landsat 5 Thematic Mapper (TM) and Landsat 8 Operational Land Imager (OLI) satellite images accessed on the United States Geological Survey (USGS) website. The Landsat scenes (path-row) used were 05-66, 06-66 and 06-65 from the years 1988 (1 year before the official creation of the SDNP), 2003 (15 years after the creation of the SDNP and half of the 30year analysis period) and 2018 (the end of the 30year analysis period and the current situation; Supplementary Table S1, available online). Image dates were chosen with the lowest possible cloud cover (<20%) during June to September.

To map land cover, we used the MaxVer supervised classification in *ArcGIS* (Esri, Redlands, CA, USA) software, which is based on the weighting of the distances between averages of the digital levels of the classes based on samples. The red, near-infrared and shortwave infrared bands were used, fused with the panchromatic band, applying the Gram–Schmidt technique to obtain a spatial resolution of 15 m. This merger of bands was used only for the year 2018, given that Landsat 5 TM lacks the panchromatic and two of the infrared bands, and exclusively for the purpose of viewing and adjusting the results of the supervised classification.

The land-use and land-cover classes considered in this study are adapted from the studies by Almeida et al. (2016) and Instituto Brasileiro de Geografia e Estatística (2013). The classes were: deforestation (D: deforested areas that occurred in the year of the classified image, including only clearing of primary forest, excluding secondary vegetation), mixed uses (MU: areas characterized by multiple land covers; e.g., family farming practiced in conjunction with traditional livestock raising), primary vegetation (PV: forests without significant alteration of their original structure), secondary vegetation (SV: vegetation from natural succession processes after total or partial suppression of primary vegetation by either anthropogenic actions or natural causes), pasture (P: vegetation planted for the grazing of cattle consisting of perennial forages), water bodies (WB: accumulations of water, such as rivers, streams and ponds), others (O: areas not covered by other categories, such as rock or mountain outcrops, sandbanks and roads) and unobserved areas (UA: areas with clouds or cloud shadows at the time of satellite passage).

#### Analysis

#### Accuracy of land-use mapping

The accuracy of mapping land-use change was based on the Kappa index, which estimates agreement between mapping and field observations (Meneses & Almeida 2012). The reference data used to assess the degree of accuracy of the results were obtained by field sampling with a drone (DJI Remotely Piloted Aircraft, Dà-Jiāng Innovations Science and Technology Co., Shenzhen, China). Due to the challenging accessibility and remoteness of the SDNP, which is the westernmost park in Brazil, the 413 points were collected and georeferenced where the Juruá River forms *c*. 200 km of the park border using a Global Positioning System receiver from 6 to 14 November 2018 (Supplementary Fig. S1).

#### Land-use patterns

We analysed the land-use patterns for the years 1988, 2003 and 2018 in the SDNP and within a 5-km buffer of the park's border (buffer zone defined to capture the influence of the SDNP through its main access road and the sites of human occupation along the banks of the Juruá River), excluding the border with Peru. A 5-km buffer zone was chosen for the analysis to capture the area most proximate, and thus most similar, to the interior of the park. The area of each land-use class was calculated for all years analysed inside and outside of the SDNP. The percentage losses of forest cover inside and outside of the SDNP were calculated in order to understand the effectiveness of the SDNP at maintaining forest. We built a land-use/cover transition matrix over the three decades in order to understand the main replacements of forest cover over time.

#### Results

#### Mapping assessment

Although 1% of the forest area had been lost prior to 1988, between 1988 and 2018 the SDNP lost only an additional 1% of its native forest cover (Fig. 2 & Table 1). The evaluation of the supervised classification for the year 2018 indicated an overall accuracy of 91.5% in relation to the points collected in the field (of the total

of 413 samples, 378 were correctly classified) and the Kappa index reached 89.3% (≥81% indicates good agreement; Table S2).

The areas occupied by pasture and deforestation account for 15–17% of the non-forest area, showing the stability of human occupation. The main change in land cover in the SDNP was the transition from PV to SV classes and MU, representing 67–71% of the area without native forest (1447 ha), with 43–45% being SV.

Forest loss within the buffer zone was greater than inside the SDNP (Fig. 2), with the buffer zone experiencing a 10% loss of forest over the 30 years (Table 2). The replacement of PV by the P and D classes represented 22% of the changes in the PV class between 1988 and 2003 and 39% of the changes in the most recent period (2003–2018). The abandonment of agriculture and ranching to SV and subsequent transition to 'forest' (secondary forest with a structure similar to that of PV) was 45% during 1988–2003, decreasing to 25% during 2003–2018. The maintenance over time of productive areas (P and MU) was greater in the 5-km buffer zone than in the Park. Areas and percentages of land-use and land-cover classes in 1988, 2003 and 2018 in the SDNP and in the 5-km buffer zone are given in the Supplementary Material (Tables S3 & S4).

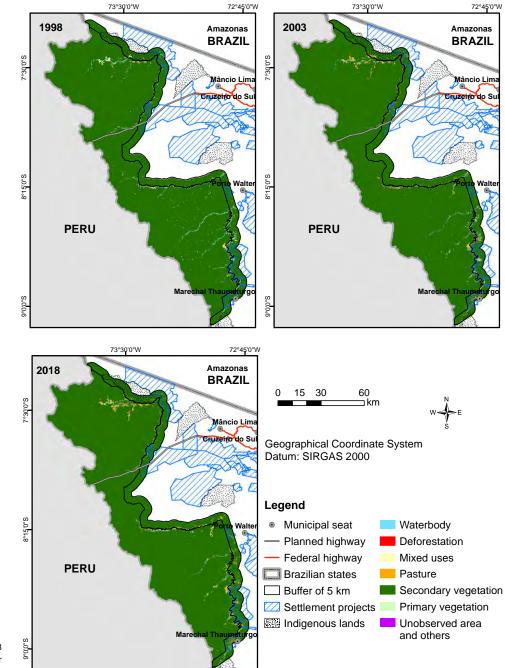
The natural land covers in the SDNP, such as WB along the Juruá and Moa rivers, occupied an average of 1601 ha. The transitions observed between classes O and W for PV and SV in the matrices (Tables 1 & 2) are due to the changes that occurred in the courses of the Juruá, Juruá Mirim, Moa and other rivers because they meander over time. The O class includes areas disturbed by natural landslides in the Jaquirana, Moa, Juruá Mirim and Rio Branco mountains, caused periodically by excessive rain or by earthquakes.

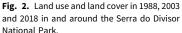
#### Discussion

#### Land-use change in and around the Serra do Divisor National Park

Researchers have shown that the establishment of PAs mitigates deforestation in the Amazon (Barber et al. 2014, Schleicher et al. 2017). Our results reveal that the SDNP has been effective in controlling the expansion of deforestation, as in 30 years the park lost only 1% of its forest cover, in contrast with its immediate surroundings, which lost 10% of their forest cover. Transitions from forest to other uses occurred in settlement projects to the south of the SDNP, such as the Amônia Settlement Project (Fig. 3), consistent with similar situations identified in the state of Pará (Calandino et al. 2012) and the state of Roraima (Barni et al. 2015). The production of manioc (cassava) flour along with other agricultural crops is related to the expansion of the areas of mixed uses in the SDNP (455 ha in 1988 increasing to 2972 ha in 2018) and in its surroundings (340 ha in 1988 increasing to 5282 ha in 2018). Manioc flour is the most important regional product and is the main source of income for most residents of the study region.

The consistent monitoring of land-cover and land-use transitions indicates that some administrative territorial designations are more effective at deterring deforestation, associated with inspection and control policies being more effective at reducing deforestation (Hargrave & Kis-Katos 2013). That said, CUs such as national parks have increasingly struggled to protect standing forest under the pro-agribusiness and developmentalist rhetoric and policies of Brazilian President Bolsonaro (Fearnside et al. 2020, Ferrante & Fearnside 2021). In addition, we recognize that





the SDNP's remoteness, borderland location, lack of roads and mountainous interior may make deforestation more difficult and bias our results (Joppa & Pfaff 2009). Similarly, while navigable rivers provide axes of deforestation in the Amazon, less deforestation takes place within PAs bordering rivers (Barber et al. 2014), thus potentially increasing negative spatial spill-over of deforestation on the riverside outside of PAs and other administrative units (Pfaff & Robalino 2017, Pfaff et al. 2007). This research also does not analyse the potentially important role of crop (e.g., manioc) and livestock (e.g., cattle) prices or changing conservation policies on deforestation within or outside of the Park (Assunção et al. 2015). However, a new international highway through the SDNP would provide access to and accelerate deforestation in previously remote, borderland and mountainous areas while also likely increasing spill-over via secondary roads and providing opportunities for more price- and policy-induced deforestation along roadsides.

## Threats to the Serra do Divisor National Park

The SDNP is considered to have high biodiversity and endemism, as well being a centre for ecotourism (Scarcello & Passos 1998). The SDNP is recognized for the wealth of mammal and amphibian species (Bernarde et al. 2016) and for its endemic species, such as *Thamnophilus divisorius* (Whitney et al. 2004) and various hymenopterans (Azevedo & Batista 2002) and lepidopterans (Dolibaina et al. 2015). The isolated low mountains of the SDNP and its Peruvian sister park are geologically distinct from Table 1. Transition matrices of land use and land cover for the years 1988, 2003 and 2018 and the transitions between classes in the periods 1988–2003 and 2003–2018 within the Serra do Divisor National Park.

			Area in 2003 (ha)							
	Thematic class	PV	Р	D	SV	W	0	MU	UA	Total
Area in 1988 (ha)	PV	822 956	719	375	3157	457	636	1268	10	829 578
	Р	109	699	<1	208	6	15	60	0	1098
	D	85	22	<1	82	3	5	31	0	228
	SV	1475	453	7	1138	38	63	263	0	3436
	W	349	7	2	39	788	407	22	0	1614
	0	419	17	1	109	217	398	29	0	1190
	MU	177	25	1	178	9	18	48	0	455
	UA	0	0	0	0	0	0	0	0	-
	Total	825 570	1942	387	4911	1517	1542	1720	10	837 599
Area in 2018 (ha)	PV	818 319	98	152	2202	285	612	479	3	822 150
	Р	795	1438	49	484	8	24	286	0	3084
	D	275	0	1	6	1	4	3	0	290
	SV	3260	287	125	1581	29	119	560	2	5963
	W	528	4	4	49	787	364	17	<1	1754
	0	522	9	5	71	383	364	23	<1	1377
	MU	1865	106	51	516	23	55	351	5	2972
	UA	6	0	0	2	<1	0	1	0	10
	Total	825 570	1942	387	4911	1517	1542	1720	10	837 599

D = deforestation; MU = mixed uses; O = others; P = pasture; PV = primary vegetation; SV = secondary vegetation; UA = unobserved area; W = waterbodies.

 Table 2.
 Transition matrices of land use and land cover for the years 1988, 2003 and 2018 and the transitions between classes in the periods 1988–2003 and 2003–2018 in the buffer zone of the Serra do Divisor National Park.

			Area in 2003 (ha)							
	Thematic class	PV	Р	D	SV	W	0	MU	UA	Total
Area in 1988 (ha)	PV	174 511	1457	327	3798	562	471	1396	67	182 588
	Р	32	171	0	95	9	15	43	0	363
	D	45	19	0	38	10	8	15	0	134
	SV	617	213	13	607	36	50	212	2	1750
	W	216	7	<1	35	1135	403	15	0	1811
	0	356	15	3	133	212	331	29	0	1078
	MU	86	43	<1	116	15	22	58	<1	340
	UA	0	0	0	0	0	0	0	0	0
	Total	175 862	1924	344	4822	1978	1299	1766	69	188 064
Area in 2018 (ha)	PV	162 019	68	106	1226	309	455	344	10	164 537
	Р	4480	1435	82	1447	14	48	481	37	8024
	D	926	3	3	17	4	7	4	0	962
	SV	3924	195	100	1191	75	168	361	10	6024
	W	377	26	2	73	974	250	20	1	1723
	0	386	42	1	75	562	283	24	3	1376
	MU	3639	153	50	777	40	87	528	8	5282
	UA	111	2	0	16	<1	1	5	0	136
	Total	175 862	1924	344	4822	1978	1299	1766	69	188 064

D = deforestation; MU = mixed uses; O = others; P = pasture; PV = primary vegetation; SV = secondary vegetation; UA = unobserved area; W = waterbodies.

most of the Amazon lowlands and provide unique landscapes and habitats in the Brazilian Amazon (Vriesendorp et al. 2006).

Since 2010, discussions have been underway between Brazil and Peru and a protocol of intentions has been signed on the construction of a highway connecting Cruzeiro do Sul (Brazil) and Pucallpa (Peru) (COSIPLAN 2017). This proposal is based on expanding the flow of Brazilian products to international markets through Peru's Pacific Ocean ports and the flow of Peruvian goods to Brazilian markets. The execution of this infrastructure project poses a great challenge to the central objective of the SDNP. Multiple studies have shown increased anthropogenic pressures after opening roads in or near CUs (Yanai et al. 2012, Vilela et al. 2020, Nascimento et al. 2021). Walker et al. (2019) and Lovejoy and Nobre (2018) have shown that deforestation associated with such infrastructure could push the Amazon rainforest beyond a 'tipping point', causing its collapse in the southern and south-western regions.

Land invasion, real-estate speculation and deforestation quickly escape government control (Ferrante & Fearnside 2020). Habitat fragmentation, forest fires and excessive hunting, amongst other forms of environmental degradation, have irreversible impacts (Laurance et al. 2009, 2014, Harrison 2011). The biodiversity of the SDNP region could be harmed due to fragmentation of the forests with the construction of the road (González-Suárez et al. 2018). Examples can be found of various highways that pass inside PAs, affecting them both directly and indirectly (Bager et al. 2015). The construction of the proposed Cruzeiro do Sul–Pucallpa road can be expected to increase social vulnerability by facilitating the trafficking of drugs, weapons, bushmeat, wild animals and natural resources (Young 2004, Suárez et al. 2009). Drug trafficking is

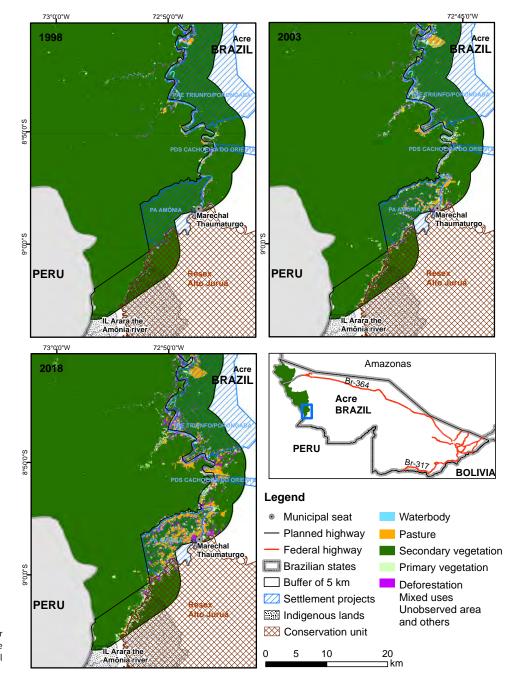


Fig. 3. Changes in land use and land cover between 1988 and 2018 in and around the Serra do Divisor National Park, Marechal Thaumaturgo region.

consolidated in the region, since Peru is one of the world's leading producers and exporters of coca derivatives, with the Peruvian department of Ucayali, of which Pucallpa is the capital, being a growing centre of such production, and Brazil is the main destination country in South America for cocaine and a key transit country for drugs headed to Europe and Africa (Santos 2010, Salisbury & Fagan 2013). The Isconahua people (an Indigenous group in a state of isolation and initial contact) would be particularly vulnerable to the road and subsequent threats such as disease, exploitation and cultural change (Krokoszynski et al. 2007).

The economic benefit used as justification for this highway, which is part of the IOC, is questionable, particularly in view of the existence of the IOS in eastern Acre, which has linked the capital of Acre (Rio Branco) with Pacific ports in Peru since the late 2000s. The IOS has been a disappointing mechanism for legitimate economic growth and produced an ongoing corruption scandal in Peru that has implicated four of the country's presidents (Alberti & Pereyra 2018, Perz & Rojas 2020). Indeed, the IOS obviates the argument that the Cruzeiro do Sul–Pucallpa highway is needed for the integration of Brazil's agricultural heartland with Peruvian ports, especially since the existing route is underutilized (Oliveira et al. 2019).

Economic cost-benefit estimates for the Cruzeiro do Sul-Pucallpa highway are negative, even without including the social and environmental costs (Glave et al. 2012, Hopkins et al. 2015). The road project also shows a persistent lack of coordination: the proposed Brazilian route would arrive at the border 25 km north of the proposed Peruvian route (Salisbury et al. 2013). The socio-environmental impacts in Acre of the proposed road need to be analysed, including estimating the losses of natural capital and ecosystem services, and potential mitigation needs to be assessed, as was done for impacts in the department of Ucayali, Peru (Mandle et al. 2013). The costs also need to be compared to alternative proposals for regional development. Measures that can reduce social and environmental costs in the face of climate change and its long-term impacts include payment for forest ecosystem services, landscape restoration and the adoption of environmentally sound agricultural practices (Lapola et al. 2018).

Due to the 'full protection' category of the SDNP in Brazil's National System of Conservation Units (SNUC), approval of the road has faced barriers due to the need to follow strict rules under Brazilian law. This protection status is threatened by the proposed downgrading to an APA (Bill 6024/2019). Some interest groups in Brazilian society claim that the social and environmental gains generated by fully protected CUs in the Amazon region restrict local economic growth, a claim that has been contested by Kauano et al. (2020).

Downgrading, downsizing and degazetting PAs are occurring worldwide and especially in Brazil (Mascia & Pailler 2011, Pack et al. 2016). Currently threatened PAs in Brazil include the Nascentes do Lago Jari National Park on Highway BR-319 (Manaus–Porto Velho), through which a road (AM-366) is planned (Fearnside et al. 2020), the Jamanxim National Forest on Highway BR-163 (Santarém–Cuiabá), through which a planned railway would pass (Chagas 2017), and the Iguaçu National Park that protects the largest remaining fragment of inland Atlantic rainforest, though which the proposed 'Caminho do Colono' road would pass (Prasniewski et al. 2020).

Motivations for downgrading PAs vary widely, but they invariably focus on access to and use of natural resources (Mascia & Pailler 2011) and may be related to the size of the PAs and local population densities (Symes et al. 2016). Re-categorization, disaffection or reduction of the limits of PAs can result in dramatically higher deforestation rates and forest carbon emissions (Forrest et al. 2015). Prior to re-categorization it is essential to carry out technical studies and public consultation on such measures (Zamadei et al. 2019). Such studies also must account for the SDNP's location and role in a transboundary conservation corridor that stretches from the Amazon River to the Madre de Dios River (Vriesendorp et al. 2006, Grupo Geográfico Transfronterizo de la Amazonía Sud Occidental 2013).

#### Implications for conservation

The SDNP still conserves *c*. 98% of its primary forest, and 48% of the total that was deforested by 2018 now has secondary vegetation. Forests in PAs store and capture atmospheric carbon, and their preservation could be an important contributor to mitigating global warming (Maxwell et al. 2019). As of 2014, the SDNP's forests held 134.9 million Mg of carbon (above and below ground), having lost only 2% of its original stock to deforestation since *c*. 1970 (Nogueira et al. 2018).

The preservation of the SDNP contributes to the food security of the inhabitants of the PA through the protein coming from wild animals (Sarti et al. 2015, Lemos et al. 2018). There are 407 families in the SDNP that use natural resources and engage in small-scale agriculture and livestock raising, in addition to the people who visit for tourism and for scientific or educational purposes. Increased involvement of the local population should be sought in the decision-making processes in PAs because this is a prerequisite for both conservation and socioeconomic development (Oldekop et al. 2016).

The south-western part of the Amazon basin, in which the SDNP is located, is an important direct and intermediate source of moisture distributed to the Prata basin that includes São Paulo and Buenos Aires (Zemp et al. 2014). Changes in land use in this region of the Amazon can weaken the forests' water recycling role, modifying precipitation thousands of kilometres from this location and intensifying climatic extremes, thus affecting agricultural productivity both inside and outside of the tropics (Lawrence & Vandecar 2015).

These impacts are in addition to the scenario investigated by Leite-Filho et al. (2020), who found that a shortening of the rainy season has occurred in southern Amazonia over a period of 15 years due to deforestation and other factors. The dry season has already lengthened in southern Amazonia (Espinoza et al. 2019). Both of these analyses could be applied to the SDNP region and its surroundings in a future study.

#### Conclusions

That the SDNP underwent much less reduction in primary forest (1%) over the 30-year period from 1988 to 2018 than did a 5-km buffer zone in the surrounding area (10%) demonstrates the importance of PAs in the Amazon region and reaffirms the roles that these areas play in maintaining forest and thus mitigating climate change and protecting biodiversity. A proposed down-grading of this national park to an APA would impact the >400 families that live in the Park, as would a proposed road that would bisect the Park. Studies show the economic infeasibility of the road project and the likely impacts on deforestation, biodiversity and the regional and continental climate regime. Stopping both the proposed road project and the downgrading of the Park's protected status would be the most appropriate way to maintain the forest in the Park and to contribute to regional and global climate stability.

Supplementary material. To view supplementary material for this article, please visit https://doi.org/10.1017/S0376892922000091.

#### Acknowledgements. None.

**Financial support.** Some of the researchers were supported by the Ministério da Ciência, Tecnologia e Inovação/Conselho Nacional de Desenvolvimento Científico e Tecnológico (grant numbers 311103/2015-4, 442650/2018-3), FINEP/Rede CLIMA (grant number 01.13.0353-00) and Ministério da Ciência, Tecnologia e Inovação/Instituto Nacional de Pesquisas da Amazônia (grant number PRJ15.125).

Conflict of interest. The authors declare none.

Ethical standards. None.

#### References

- Alberti J, Pereyra A (2018) Carretera Interoceánica IIRSA Sur de Perú: Un megaproyecto con preinversión express. Washington, DC, USA: Inter-American Development Bank.
- Almeida CA, Coutinho AC, Esquerdo JCDM, Adami M, Venturieri A, Diniz CG et al. (2016) High spatial resolution land use and land cover mapping of the Brazilian Legal Amazon in 2008 using Landsat-5/TM and MODIS data. *Acta Amazonica* 46: 291–302.
- Assunção J, Gandour C, Rocha R (2015) Deforestation slowdown in the Brazilian Amazon: prices or policies? *Environment and Development Economics* 20: 697–722.



- Azevedo CO, Batista ML (2002) New species of *Apenesia* (Hymenoptera, Bethylidae) from the Parque Nacional da Serra do Divisor, Acre, Brazil. *Revista Brasileira de Entomologia* 46: 25–32.
- Bager A, Borghi CE, Secco H (2015) The influence of economics, politics, and environment on road ecology in South America. In: *Handbook of Road Ecology* (pp. 407–413). Hoboken, NJ, USA; Chichester, UK: Wiley-Blackwell.
- Barber CP, Cochrane MA, Souza CM Jr, Laurance WF (2014) Roads, deforestation, and the mitigating effect of protected areas in the Amazon. *Biological Conservation* 177: 203–209.
- Barni PE, Pereira VB, Manzi AO, Barbosa RI (2015) Deforestation and forest fires in Roraima and their relationship with phytoclimatic regions in the northern Brazilian Amazon. *Environmental Management* 55: 1124–1138.
- Bernarde PS, Miranda DB, Albuquerque S, Turci LCB (2016) Amphibia, Anura, Hemiphractidae, *Hemiphractus helioi* Sheil and Mendelson, 2001: distribution extension in the state of Acre and second record for Brazil. *Check List* 6: 491–492.
- Calandino D, Wehrmann M, Koblitz R (2012) Contribuição dos assentamentos rurais no desmatamento da Amazônia: um olhar sobre o Estado do Pará. *Desenvolvimento e Meio Ambiente* 26: 161–170.
- Chagas VC (2017) Após veto, governo envia ao Congresso novo projeto que reduz floresta no Pará. Agência Brasil [www document]. URL https:// agenciabrasil.ebc.com.br/politica/noticia/2017-07/apos-veto-governo-enviaao-congresso-novo-projeto-que-reduz-floresta-no/
- COSIPLAN (2017) Interconexión Terrestre Pucallpa Cruzeiro do Sul. Consejo Suramericano de Infraestructura y Planeamiento de UNASUR. Washington, DC, USA: COSIPLAN.
- Dolibaina DR, Dias FMS, Mielke OHH, Casagrande MM (2015) *Argyrogrammana* Strand (Lepidoptera: Riodinidae) from Parque Nacional da Serra do Divisor, Acre, Brazil, with the description of four new species. *Zootaxa* 4028: 227–245.
- Espinoza JC, Ronchail J, Marengo JA, Segura H (2019) Contrasting north–south changes in Amazon wet-day and dry-day frequency and related atmospheric features (1981–2017). *Climate Dynamics* 52: 5413–5430.
- Fearnside PM (2021) The intrinsic value of Amazon biodiversity. *Biodiversity* and Conservation 30: 1199–1202.
- Fearnside PM, Ferrante L, Yanai AM, Isaac Júnior MA (2020) Trans-Purus: Brazil's last intact Amazon forest at immediate risk (commentary). Mongabay [www document]. URL https://news.mongabay.com/2020/11/ trans-purus-brazils-last-intact-amazon-forest-at-immediate-risk-commentary/
- Ferrante L, Fearnside PM (2020) The Amazon's road to deforestation. *Science* 369: 634.
- Fleury LC, Almeida J (2013) A construção da Usina Hidrelétrica de Belo Monte: conflito ambiental e o dilema do desenvolvimento. Ambiente & Sociedade 16: 141–156.
- Forrest JL, Mascia MB, Pailler S, Abidin SZ, Araujo MD, Krithivasan R, Riveros JC (2015) Tropical deforestation and carbon emissions from protected area downgrading, downsizing, and degazettement (PADDD). *Conservation Letters* 8: 153–161.
- Glave M, Hopkins A, Malky A, Fleck L (2012) Análisis económico de la carretera Pucallpa – Curzeiro do Sul. Rio de Janeiro, Brazil: Conservation Strategy Fund.
- González-Suárez M, Ferreira FZ, Grilo C (2018) Spatial and species-level predictions of road mortality risk using trait data. *Global Ecology and Biogeography* 27: 1093–1105.
- Grupo Geográfico Transfronterizo de la Amazonía Sud Occidental (2013) Mapa del Corredor Socioambiental de las Fronteras Amazónicas entre Ucayali, Madre de Dios (Perú), Acre (Brasil) y Pando (Bolivia). Rio Branco, Brazil: IPGH [www document]. URL https://scholarship.richmond.edu/ geography-maps/5/
- Hargrave J, Kis-Katos K (2013) Economic causes of deforestation in the Brazilian Amazon: a panel data analysis for the 2000s. *Environmental and Resource Economics* 54: 471–494.
- Harrison RD (2011) Emptying the forest: hunting and the extirpation of wildlife from tropical nature reserves. *BioScience* 61: 919–924.
- Hopkins A, Malky A, Glave M, Ventocilla R, Ledezma JC, Arana A (2015) Análisis económico y socioambiental de los proyectos de interconexión Pucallpa-Cruzeiro do Sul. Rio de Janeiro, Brazil: Conservation Strategy Fund.

- Instituto Brasileiro de Geografia e Estatística (2013) Manual técnico de uso da terra. Rio de Janeiro, RJ, Brazil: Instituto Brasileiro de Geografia e Estatística (IBGE) [www document]. URL https://biblioteca.ibge.gov.br/visualizacao/ livros/liv81615.pdf
- Jenkins CN, Joppa L (2009) Expansion of the global terrestrial protected area system. *Biological Conservation* 142: 2166–2174.
- Joppa LN, Pfaff A (2009) High and far: biases in the location of protected areas. *PLoS ONE* 4: e8273.
- Kauano EE, Silva JMC, Diniz Filho JAF, Michalski F (2020) Do protected areas hamper economic development of the Amazon region? An analysis of the relationship between protected areas and the economic growth of Brazilian Amazon municipalities. *Land Use Policy* 92: 104473.
- Killeen TJ (2007) A Perfect Storm in the Amazon Wilderness: Development and Conservation in the Context of the Initiative for the Integration of the Regional Infrastructure of South America (IIRSA). Arlington County, VA, USA: Conservation International.
- Krokoszynski L, Stoinska-Kairska I, Martyniak A (2007) Indígenas aislados en la Sierra del Divisor (Zona fronteriza Perú–Brasil). Iquitos, Lima and Poznan, Peru: UAM-AIDESEP.
- Lapola DM, Pinho P, Quesada CA, Strassburg BBN, Rammig A, Kruijt B et al. (2018) Limiting the high impacts of Amazon forest dieback with no-regrets science and policy action. *Proceedings of the National Academy of Sciences of the United States of America* 115: 11671–11679.
- Laurance WF, Clements GR, Sloan S, O'Connell CS, Mueller ND, Goosem M et al. (2014) A global strategy for road building. *Nature* 513: 229–232.
- Laurance WF, Goosem M, Laurance SGW (2009) Impacts of roads and linear clearings on tropical forests. *Trends in Ecology & Evolution* 24: 659–669.
- Lawrence D, Vandecar K (2015) Effects of tropical deforestation on climate and agriculture. *Nature Climate Change* 5: 27–36.
- Leite-Filho AT, Costa MH, Fu R (2020) The southern Amazon rainy season: the role of deforestation and its interactions with large-scale mechanisms. *International Journal of Climatology* 40: 2328–2341.
- Lemos LP, Bizri HRE, Amaral JV, Santos AS, Koga DM, Silva FE (2018) Caça de vertebrados no Parque Nacional da Serra do Divisor, Acre. *Biodiversidade Brasileira* 8: 69–88.
- Lovejoy TE, Nobre C (2018) Amazon tipping point. Science Advances 4: eaat2340.
- Mandle L, Tallis H, Vogl AL, Wolny S, Touval J, Sotomayor L et al. (2013) Can the Pucallpa–Cruzeiro do Sul road be developed with no net loss of natural capital in Peru? A framework for including natural capital in mitigation [www document]. URL https://naturalcapitalproject.stanford.edu/sites/g/ files/sbiybj9321/f/publications/can-the-pucallpa-cruzeira-do-sul-road-bedeveloped-with-no-net-loss-of-natural-capital-in-peru.pdf
- Mascia MB, Pailler S (2011) Protected area downgrading, downsizing, and degazettement (PADDD) and its conservation implications. *Conservation Letters* 4: 9–20.
- Maxwell SL, Evans T, Watson JEM, Morel A, Grantham H, Duncan A et al. (2019) Degradation and forgone removals increase the carbon impact of intact forest loss by 626%. *Science Advances* 5: eaax2546.
- Meneses PR, Almeida T (2012) Introdução ao processamento de imagens de sensoriamento remoto. Brasília, Brazil: Universidade de Brasília.
- Nascimento ES, Silva SS, Bordignon L, Melo AWF, Brandão A, Souza CM, Silva Junior CHL (2021) Roads in the southwestern Amazon, State of Acre, between 2007 and 2019. *Land* 10: 106.
- Nelson A, Chomitz KM (2011) Effectiveness of strict vs. multiple use protected areas in reducing tropical forest fires: a global analysis using matching methods. *PLoS ONE* 6: e22722.
- Nogueira EM, Yanai AM, Vasconcelos SS, Graça PMLA, Fearnside PM (2018) Carbon stocks and losses to deforestation in protected areas in Brazilian Amazonia. *Regional Environmental Change* 18: 261–270.
- Oldekop JA, Holmes G, Harris WE, Evans KL (2016) A global assessment of the social and conservation outcomes of protected areas. *Conservation Biology* 30: 133–141.
- Oliveira AS, Soares-Filho, BS, Costa MA, Lima L, Garcia RA, Rajão R, Carvalho-Ribeiro SM (2019) Bringing economic development for whom? An exploratory study of the impact of the Interoceanic Highway on the livelihood of smallholders in the Amazon. *Landscape and Urban Planning* 188: 171–179.

- Pack SM, Ferreira MN, Krithivasan R, Murrow J, Bernard E, Mascia MB (2016) Protected area downgrading, downsizing, and degazettement (PADDD) in the Amazon. *Biological Conservation* 197: 32–39.
- Perz S, Rojas RO (2020) Do infrastructure improvements for regional integration increase traffic volume? The case of the Inter-Oceanic Highway in the southwestern Amazon. *Journal of Latin American Geography* 19: 243–264.
- Pfaff A, Robalino J (2017). Spillovers from conservation programs. Annual Review of Resource Economics 9: 299–315.
- Pfaff A, Robalino J, Walker RT, Aldrich S, Caldas M, Reis E et al. (2007) Road investments, spatial spillovers, and deforestation in the Brazilian Amazon. *Journal of Regional Science* 47: 109–123.
- Prasniewski VM, Szinwelski N, Sobral-Souza T, Kuczach A, Brocardo CR, Sperber CF, Fearnside PM (2020) Parks under attack: Brazil's Iguaçu National Park illustrates a global threat to biodiversity. *Ambio* 49: 2061–2067.
- Ruaro, R, Laurance WF (2022) Pending bill could devastate Brazil's Serra do Divisor National Park. *Nature Ecology & Evolution* 6: 120–121.
- Salisbury DS, Fagan C (2013). Coca and conservation: cultivation, eradication, and trafficking in the Amazon borderlands. *Geojournal* 78: 41–60.
- Salisbury DS, Moreno MCS, Torres LD, Vásquez RG, Diaz JS, Tipula P et al. (2013) Border integrations: the fusion of political ecology and land-change science to inform and contest transboundary integration in Amazonia. In: Land Change Science and Political Ecology: Synergies and Divergences (pp. 129–149). London, UK: Earthscan.
- Santos M (2010) Peru: cultivo de coca, cocaína e combate ao narcotráfico. *Meridiano* 47: 14.
- Sarti F, Adams C, Morsello C, van Vliet N, Schor T, Yagüe B et al. (2015) Beyond protein intake: bushmeat as source of micronutrients in the Amazon. *Ecology and Society* 20: 22.
- Scarcello M, Passos V (1998) Plano de Manejo do Parque Nacional da Serra do Divisor [www document]. URL https://www.icmbio.gov.br/portal/images/ stories/docs-planos-de-manejo/parna\_serra\_divisor\_pm.pdf
- Schleicher J, Peres CA, Amano T, Llactayo W, Leader-Williams N (2017) Conservation performance of different conservation governance regimes in the Peruvian Amazon. *Scientific Reports* 7: 11318.
- Silveira M, Daly DC, Salimon CI, Wadt PGS, Amaral EF, Pereira MG, Passos V (2008) Ambientes físicos e coberturas vegetais do Acre. In: *Primeiro Catálogo da flora do Acre, Brasil* (pp. 36–63). Rio Branco, Brazil: Universidade Federal do Acre.

- Soares-Filho B, Moutinho P, Nepstad D, Anderson A, Rodrigues H, Garcia R et al. (2010) Role of Brazilian Amazon protected areas in climate change mitigation. Proceedings of the National Academy of Sciences of the United States of America 107: 10821–10826.
- Suárez E, Morales M, Cueva R, Utreras B, Zapata-Ríos G, Toral E et al. (2009) Oil industry, wild meat trade and roads: Indirect effects of oil extraction activities in a protected area in north-eastern Ecuador. *Animal Conservation* 12: 364–373.
- Symes WS, Rao M, Mascia MB, Carrasco LR (2016) Why do we lose protected areas? Factors influencing protected area downgrading, downsizing and degazettement in the tropics and subtropics. *Global Change Biology* 22: 656–665.
- Vilela T, Harb AM, Bruner A, Arruda VLS, Ribeiro V, Alencar AAC et al. (2020) A better Amazon road network for people and the environment. *Proceedings* of the National Academy of Sciences of the United States of America 117: 7095–7102.
- Vriesendorp C, Schulenberg T, Alverson WS, Moskovits DK, Moscoso JR (2006) Perú: Sierra del Divisor. Chicago, IL, USA: The Field Museum.
- Walker RT, Simmons C, Arima E, Galvan-Miyoshi Y, Antunes A, Waylen M, Irigaray M (2019) Avoiding Amazonian catastrophes: prospects for conservation in the 21st century. One Earth 1: 202–215.
- Whitney BM, Oren DC, Brumfield RT (2004) A new species of *Thamnophilus antshrike* (Aves: Thamnophilidae) from the Serra do Divisor, Acre, Brazil. *The Auk* 121: 1031–1039.
- Wiens JA, Seavy NE, Jongsomjit D (2011) Protected areas in climate space: what will the future bring? *Biological Conservation* 144: 2119–2125.
- Yanai AM, Fearnside PM, Graça PMLA, Nogueira EM (2012) Avoided deforestation in Brazilian Amazonia: simulating the effect of the Juma Sustainable Development Reserve. *Forest Ecology and Management* 282: 78–91.
- Young KR (2004) Environmental and social consequences of coca/cocaine in Peru: policy alternatives and a research agenda. In: MK Steinberg, JJ Hobbs, K Mathewson (eds.), *Dangerous Harvest: Drug Plants and the Transformation of indigenous Landscapes* (pp. 249–273). Oxford, UK: Oxford University Press.
- Zamadei T, Heimann JP, Pires PTL (2019) Recategorização de unidades de conservação: estudo de caso Reserva Biológica Nascentes da Serra do Cachimbo – PA, Brasil. *Ciência Florestal* 29: 1796–1808.
- Zemp DC, Schleussner CF, Barbosa HMJ, van der Ent RJ, Donges JF, Heinke J et al. (2014) On the importance of cascading moisture recycling in South America. *Atmospheric Chemistry & Physics* 14: 13337–13359.

# **Supplementary Material**

# Serra do Divisor National Park: a protected area under threat in the

# southwestern Brazilian Amazon

Diogo Mitsuru Koga<sup>1</sup>, Irving Foster Brown<sup>2,4\*</sup>, Philip Martin Fearnside<sup>1</sup>, David S.

Salisbury<sup>3</sup>, Sonaira Souza da Silva<sup>4</sup>

<sup>1</sup>Instituto Nacional de Pesquisas da Amazônia, Programa de Pós-Graduação em Gestão de Áreas

Protegidas na Amazônia, Manaus, AM, Brasil, 69067375, diogomkoga@hotmail.com;

pmfearn@inpa.gov.br

<sup>2</sup>Woodwell Climate Research Center, 149 Woods Hole Rd., Falmouth, MA 02540, USA,

fbrown@woodwellclimate.org

<sup>3</sup>University of Richmond, Department of Geography and the Environment, Westhampton Way,

Richmond, VA, 23173, USA, dsalisbu@richmond.edu

<sup>4</sup>Universidade Federal do Acre, Geoprocessing Laboratory Applied to the Environment - LabGAMA,

Cruzeiro do Sul, AC, CEP 69.980-000, Brazil, sonaira.silva@ufac.br,

\* Corresponding author

Table S1. Satellite imagery used.	. 2
<b>Table S2.</b> Confusion matrix based on the 2018 image. The main diagonal (values in bold)         corresponds to correctly sorted samples.	. 2
<b>Table S3.</b> Area in hectares and percentage of land-use and land-cover classes in the SDNP in 1988, 2003 and 2018.	3
<b>Table S4</b> . Area in hectares and percentage of land-use and land-cover classes in         the area surrounding the SDNP in 1988, 2003 and 2018	3
Figure S1. Location of samples (aerial photos) collected with the DJI Spark drone	4

Year	Path/Row	Sensor/Satellite	Scene date	Spatial Resolution
	05/66	TM/Landsat 5	13/06/1988	30 m
1988	06/66	TM/Landsat 5	07/08/1988	30 m
	06/65	TM/Landsat 5	07/08/1988	30 m
	05/66	TM/Landsat 5	23/06/2003	30 m
2003	06/66	TM/Landsat 5	02/09/2003	30 m
	06/65	TM/Landsat 5	02/09/2003	30 m
	05/66	OLI/Landsat 8	19/08/2018	30 m
2018	06/66	OLI/Landsat 8	27/09/2018	30 m
	06/65	OLI/Landsat 8	11/09/2018	30 m

 Table S1. Satellite imagery used.

**Table S2.** Confusion matrix based on the 2018 image. The main diagonal (values in bold) corresponds to correctly sorted samples.

		Field truth						
Thematic class	PV	Р	D	SV	W	0	MU	TOTAL
Primary vegetation (PV)	126			15		2	2	145
Pasture (P)		50		1			2	53
Deforestation (D)			18					18
Secondary vegetation (SV)				59			3	62
Waterbodies (W)	1				23			24
Others (O)						25	1	26
Mixed uses (MU)		3		3		2	77	85
TOTAL	127	53	18	78	23	29	85	413

	Interior of the SDNP								
Thematic class / Year	198	8	200	3	2018				
Thematic class / Year	Area	%	Area	%	Area	%			
Primary vegetation (PV)	829,578	99.04	825,570	98.56	822,150	98.16			
Pasture (P)	1,098	0.13	1,942	0.23	3,084	0.37			
Deforestation (D)	228	0.03	387	0.05	290	0.03			
Secondary vegetation (SV)	3,436	0.41	4,911	0.59	5,963	0.71			
Waterbodies (W)	1,614	0.19	1,517	0.18	1,754	0.21			
Others (O)	1,190	0.14	1,542	0.18	1,377	0.16			
Mixed uses (MU)	455	0.05	1,720	0.21	2,972	0.35			
Unobserved area (UA)	0	0.00	10	0.00	9	0.00			
Total	837,599	100	837,599	100	837,599	100			

 Table S3. Area in hectares and percentage of land-use and land-cover classes in the SDNP in 1988, 2003 and 2018.

**Table S4**. Area in hectares and percentage of land-use and land-cover classes in the area surrounding the SDNP in 1988, 2003 and 2018.

	Surrounding area (5-km buffer) of the SDNP								
Thematic class / Year	198	8	200	3	2018				
Thematic class / Year	Area	%	Area	%	Area	%			
Primary vegetation (PV)	182,588	97.09	175,862	93.51	164,537	87.49			
Pasture (P)	363	0.19	1,924	1.02	8,024	4.27			
Deforestation (D)	134	0.07	344	0.18	962	0.51			
Secondary vegetation (SV)	1,750	0.93	4,822	2.56	6,024	3.20			
Waterbodies (W)	1,811	0.96	1,978	1.05	1,723	0.92			
Others (O)	1,078	0.57	1,299	0.69	1,376	0.73			
Mixed uses (MU)	340	0.18	1,766	0.94	5,282	2.81			
Unobserved area (UA)	0	0.00	69	0.04	136	0.07			
Total	188,064	100	188,064	100	188,064	100			

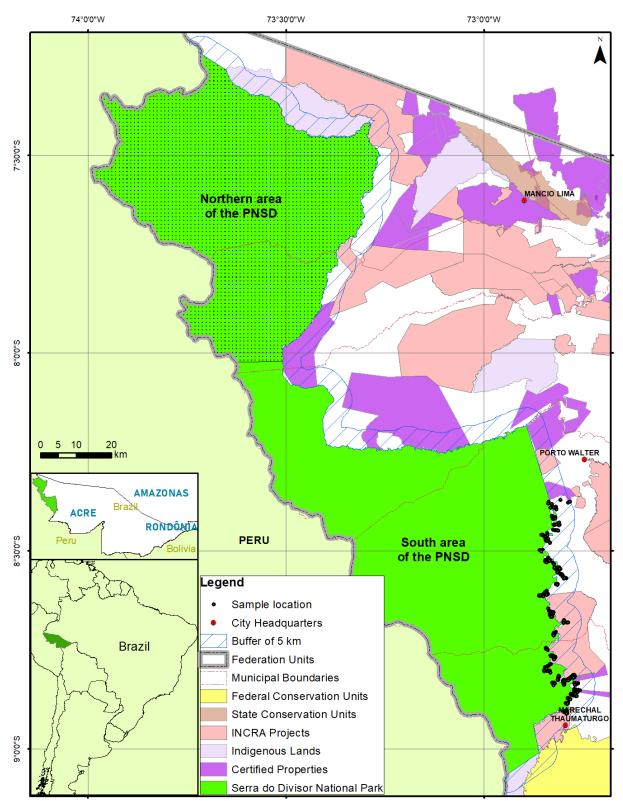


Figure S1. Location of samples (aerial photos) collected with the DJI Spark drone.

# Adjustment of mapping land-use change

During the supervised classification process, adjustments were made in the sample collection between the land-use classes. This adjustment was based on the analysis of the product of the first classification with images with better spatial resolution (15 m) obtained through the fusion of the panchromatic band with red, near-infrared and short-wave infrared bands, both Landsat 8 bands. This product allowed the reevaluation of confusing classes, such as pasture, secondary vegetation and others. From this analysis, it was possible to adjust the collection of samples and improve our supervised classification.