

This file has been cleaned of potential threats.

If you confirm that the file is coming from a trusted source, you can send the following SHA-256 hash value to your admin for the original file.

a17ce82f3dd85f09975017856f21450897b91b8ae0c820b72fa1785cce84dc38

To view the reconstructed contents, please SCROLL DOWN to next page.

The text that follows is a PREPRINT.
O texto que segue é um PREPRINT.

Please cite as:

Favor citar como:

Yanai, A.M., P.M.L.A. Graça, M.I. S.
Escada, L.G. Ziccardi & P.M. Fearnside.
2020. Deforestation dynamics in Brazil's
Amazonian settlements: Effects of land-
tenure concentration. *Journal of
Environmental Management*
<https://doi.org/10.1016/j.jenvman.2020.110555> (In press).

ISSN: 0301-4797

Copyright: Elsevier

The original publication is available at:
O trabalho original está disponível em:

<https://doi.org/10.1016/j.jenvman.2020.110555>
<https://www.journals.elsevier.com/journal-of-environmental-management>

1
2 **Title:** Deforestation dynamics in Brazil's Amazonian settlements: Effects of land-tenure
3 concentration

4
5 **Authors:** Aurora Miho Yanai¹, Paulo Maurício Lima de Alencastro Graça^{1,4}, Maria Isabel
6 Sobral Escada², Leonardo Guimarães Ziccardi^{1,3}, Philip Martin Fearnside^{1,4,5}

7
8 ¹Department of Environmental Dynamics, National Institute for Research in Amazonia (INPA),
9 Av. André Araújo n° 2936, CEP 69067-375, Manaus, Amazonas, Brazil.

10
11 ²Department of Image Processing, National Institute for Space Research (INPE), Av.
12 Astronautas, n° 1758, CEP 12227-010, São José dos Campos, São Paulo, Brazil

13
14 ³Department of Forestry, Michigan State University, East Lansing, MI, 48824, USA

15
16 ⁴Brazilian Research Network on Climate Change (RedeClima)

17
18 ⁵Corresponding author

19
20 Email addresses: yanai@inpa.gov.br; pmlag@inpa.gov.br; isabel.escada@dpi.inpe.br;

21 leonardo.g.ziccardi@gmail.com; pmfearn@inpa.gov.br

22 Tel: +55(92) 3643-1822

23

24
25 **Title:** Deforestation dynamics in Brazil's Amazonian settlements: Effects of land-tenure
26 concentration
27

28
29 **Abstract**

30 Brazil's Amazon deforestation is a major global and national environmental concern, and the
31 ability to model and project both its course and the effect of different policy options depends on
32 understanding how this process occurs at present and how it might change in the future. The
33 present paper addresses one key factor in Amazon deforestation: land-tenure concentration in
34 settlements. Brazil's policies for establishing and regulating settlement projects represent
35 critical government decisions shaping the landscape in the 5×10^6 km² Legal Amazonia region.
36 We used remote-sensing data and information provided by the National Institute for
37 Colonization and Agrarian Reform (INCRA) to evaluate the effect of land-tenure concentration
38 in a settlement project (*Projeto de Assentamento*) located in a frontier area where cattle-
39 ranching is expanding. We identified the actors and their deforestation patterns in the Matupi
40 settlement in the southern part of Brazil's state of Amazonas. We spatially identified actors who
41 concentrated "lots" (the parcels of land distributed to individual settlers) in 2011 and assessed
42 whether the concentration was done by individual landholders or by "families" (where members
43 merged their lots and the clearing was done together). Deforestation rates (1995-2011) were
44 estimated for each type of actor and the trajectory of deforestation in the settlement (cumulative
45 deforestation to 1994 and annual deforestation 1995-2016) was also analyzed. Concentrators
46 occupied 28% (9653 ha) of the settlement and 29% of the lots (152 lots) analyzed; the numbers
47 of lots concentrated ranged from two to ten. Concentrators of two lots and non-concentrators
48 were the predominant actor types in the settlement. The mean annual clearing per landholding
49 for concentrators of two lots (families: 4.1 ± 2.8 ha (mean \pm SD); individuals: 5.1 ± 4.6 ha) was
50 greater than for non-concentrators (1.7 ± 1.2 ha), despite their having similar patterns of small
51 clearings. Concentrators of three or more lots had mean annual clearing per landholding
52 between 6.2 ± 12.2 ha and 23.9 ± 38.7 ha and, the pattern of patches cleared per year > 34 ha in
53 area was predominant. The deforestation rate per lot was higher among concentrators as
54 compared to non-concentrators, showing that lot concentration speeds deforestation. Analysis of
55 deforestation patterns helps to better understand the process of lot concentration by spatially
56 identifying the predominant patterns of each type of actor. The approach used in our study could
57 assist authorities in identifying and monitoring land-tenure concentration in settlements.
58 Agrarian-reform policymakers need to monitor this process, since it speeds deforestation in
59 Amazonian settlement projects, as well as undermining the social objectives of the agrarian-
60 reform program.

61
62 **Keywords:** Agrarian reform; Settlement project; Colonization; Deforestation pattern; Amazon
63 forest; Land concentration
64

65 **Highlights:**

- 66 • Deforestation in Brazilian Amazonia is increased by land-tenure concentration.
67 • Settlers receive 1 lot per family, but newcomers buy out the original settlers.
68 • "Concentrators" in settlements establish ranches of 2-10 lots (56 to 600 ha).
69 • In the Matupi settlement, 29% of the lots had been concentrated after 16 years.
70 • Concentrators with ≥ 3 lots typically clear in patches > 34 ha in area.

72

73 **1. Introduction**

74

75

76

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

93

94

95

96

97

98

99

100

101

102

103

104

105

106

107

108

109

110

111

112

113

114

115

116

117

118

119

120

Brazil's Amazonian settlements have an important role in the region's land-use dynamics. Direct and indirect vectors of deforestation (e.g., extensive cattle ranching and illegal occupation of several lots by a single landholder) contribute to increasing deforestation rates in settlements (Alencar et al., 2016). Because most settlements are located near major roads (e.g., the Transamazon Highway), deforestation pressure in these areas tends to be intense (Godar et al., 2012a). Deforestation results in the loss of important environmental services provided by the forest, such as maintenance of water cycling, carbon stocks and biodiversity (Fearnside, 1997, 2008a).

Settlements contributed 17% (160,410 km²) of the total clearing (clearcutting of both forest and non-forest vegetation) from the "premodern" condition to 2013 in Brazil's 5 × 10⁶ km² administrative region denominated "Legal Amazonia", which represent 20% (2.6 Pg C) of the total carbon lost in Legal Amazonia through 2013 (Yanai et al., 2017). "Premodern" refers to a time prior to major increases in disturbances beginning in approximately 1970 (Nogueira et al., 2015).

"Federal settlement project" (*Projeto de Assentamento Federal*) is the type of settlement with the largest number of settlements and encompasses 72% (115,634 km²) of the total clearing in settlements (Yanai et al., 2017). Federal settlement projects are established by Brazil's National Institute for Colonization and Agrarian Reform (INCRA), which distributes plots of land called "lots" (*lotes*) with one lot for a single person or family. When a settlement begins, all or almost all lots are held by individual families (i.e., "non-concentrators"), but as time passes many original settlers sell their lots to wealthier neighbors or to newcomers who "concentrate" several lots to manage the area as a larger property, even though the lots are held under different names. When the original settlers sell their lots to wealthier newcomers, this creates a new wave of landless migrants, leading to a continued cycle of land invasion and subsequent legalization and/or resettlement in new INCRA projects (Fearnside, 2001).

In 2017, Law 13,465 (formerly MP-759), popularly known as the "land-grabbers' law" or "*lei da grilagem*," was passed allowing illegal land claims up to 2500 ha to be legalized (Brazil, PR, 2017, Art. 6). This law also specifies that illegally occupied lots in settlement projects can be legalized after only two years of occupation (Art. 26B) and that lots can be sold after 10 years of legal occupation (Arts. 18, §1 & 22, §1). In addition, the law specifies (Art. 17, §6) that settlements be considered "consolidated" 15 years after they were founded (thereby allowing lots to be sold, whether or not the same owner has occupied the lot for 10 years). A particularly pernicious effect for settlements is ending a provision that allows settlers to start paying installments owed to the government for the original purchase of the lot only after adequate infrastructure (access roads, etc.) has been installed (e.g., Branford and Torres, 2017). These debts can now be called for immediate payment, and this can be demanded independent of the adequacy of infrastructure (Art. 17, §8). All of these provisions can be expected to result in the less-wealthy settlers, who have only one lot, selling their land to wealthier neighbors or to newcomers. Irrespective of the effect of lot concentration in speeding deforestation, newcomers who buy lots in settlement areas have been found to clear forest at a substantially faster rate per lot than the original occupants (Carrero and Fearnside, 2011; Fearnside, 1987).

Land concentration is an important issue in Amazonian rural settlements because it violates the principles of Brazil's agrarian reform program, which is intended to distribute land to landless families. In addition, concentration of lots transforms settlements into large cleared areas used mainly for cattle pasture (Browder et al., 2008; Carrero and Fearnside, 2011; Martins and Pereira, 2012). For cattle ranchers, one of the main motivations for land concentration is

121 expansion of pasture. Because law enforcement is currently not sufficient to control this
122 process, concentration of lots is a typical feature of settlement projects.

123 The present study addresses the question of whether the effect of lot concentration
124 results in distinct patterns and rates of deforestation between concentrators (either families or
125 individuals) and non-concentrators. We answer the question by (1) spatially identifying
126 concentrators and non-concentrators and whether concentration is done by “individuals” (i.e.,
127 several lots identified by INCRA as occupied by a single person) or “families” (i.e., a family
128 with lots in the names of several family members), (2) defining typologies of deforestation
129 based on the types of actors, remote-sensing data and data-mining techniques and (3) evaluating
130 the rates and trajectories of deforestation through the time in each type of land-tenure
131 concentration.

132 The term “deforestation pattern” refers to a spatial configuration of patches of
133 deforestation with similarities in size, shape and location that can be mapped from satellite
134 imagery (Zipperer, 1993; Geist and Lambin, 2001; dos Santos Silva et al., 2008). The term
135 “actors” refers to landholders (either individuals or families), whether or not they were settled
136 by INCRA.

137 A spatial and temporal analysis at the level of “polygons” (areas on a digital map in a
138 geographical information system, with each polygon enclosed by a continuous perimeter and
139 associated with attributes such as land-use type) can provide data at the patch scale in order to
140 evaluate and understand changes resulting from human action in space and through time (Lu et
141 al., 2013). Identifying the actor types and the deforestation patterns associated with them can
142 improve our comprehension of how carbon stocks have been lost by the different actor
143 categories and how deforestation might proceed in the future as the process of land
144 concentration continues. Understanding the deforestation behavior of different actor types is
145 essential if the future course of land-use change is to be predicted and appropriate measures
146 taken to avoid unfavorable outcomes.

147

148 **2. Materials and Methods**

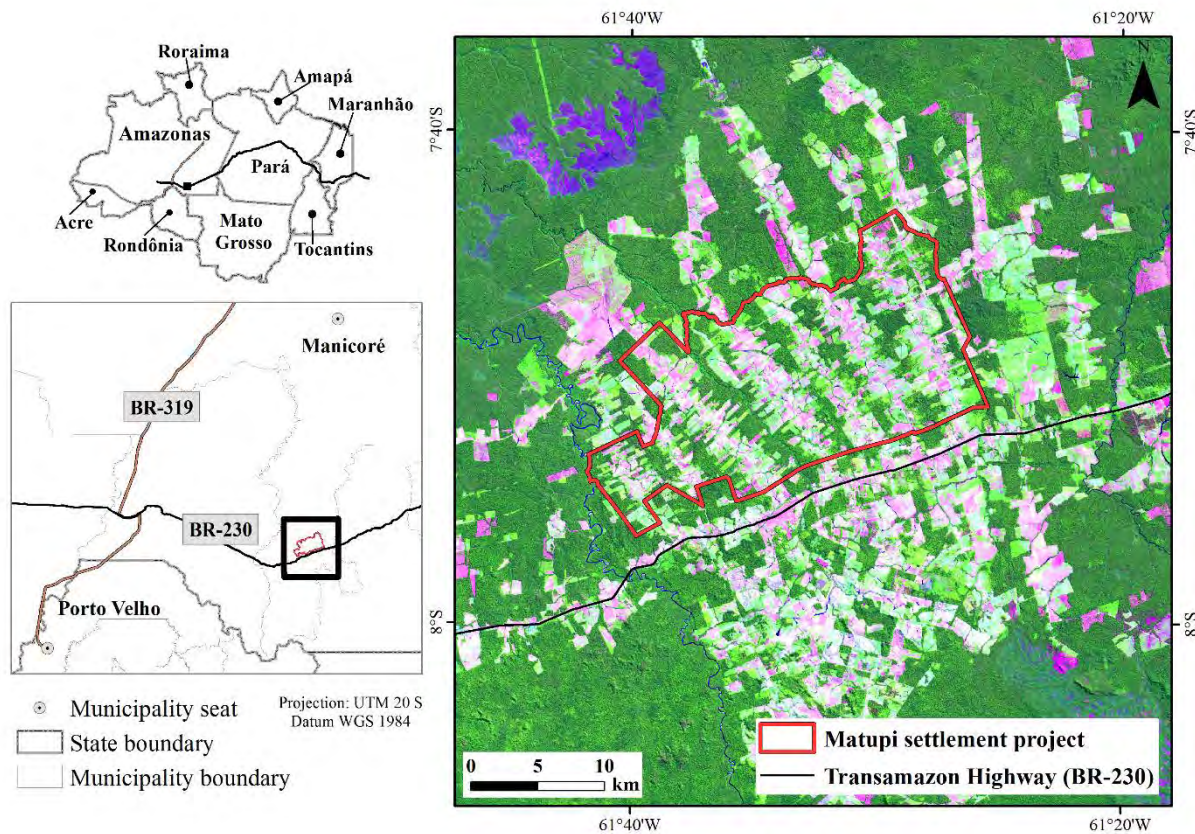
149

150 *2.1. Study area*

151

152 The present study was carried out in the Matupi Federal Settlement Project in Matupi
153 District. A “district” is an administrative unit within a municipality, in this case the municipality
154 of Manicoré in the state of Amazonas, Brazil. The Matupi settlement is located in the southern
155 part of Amazonas state near the Transamazon Highway (BR-230), which provides a road
156 connection to the state of Rondônia (a major source of migration) via the BR-319 Highway,

157 which connects Porto Velho (Rondônia) with Manaus (Amazonas) (Fig. 1).



158
159 **Fig. 1** Location of the study area. Landsat-8 OLI image (2016): R (6), G (5), B (4).
160

161 Most actors in Amazonian settlements originate from locations near the settlement or
162 from the southern and southeastern regions of Brazil (Fujisaka et al., 1996; Fearnside, 2008b;
163 Caviglia-Harris et al., 2013). Land prices are low in settlements in frontier areas as the Matupi
164 settlement, which attracts farmers from Rondônia, where the farmers in this former frontier area
165 can sell their land for a good price and use the proceeds to buy a larger area in an area where the
166 deforestation dynamic is intense.

167 Matupi District (formerly known as “km 180”) is an area characterized by expansion of
168 logging and cattle ranching. This general area was indicated as having a very high density of
169 forest loss (>10 km² per 100 km² of land area) from 2001 to 2014 (Kalamandeen et al., 2018).
170 Carbon loss in the Matupi settlement through 2013 was estimated at 3,389,406 Mg C (18,168 ha
171 of area cleared), while estimated carbon stock in the remaining forest in 2013 (16,762 ha) was
172 3,129,204 Mg C (Yanai et al., 2017).

173 The Matupi settlement was officially created on 20 July 1992, initially with 465 lots
174 covering 30,810 ha. However, the occupation process in the Matupi settlement began in 1995
175 with the establishment of 91 families (da Silva et al., 2011). In 1997 the settlement area
176 officially increased to 34,345 ha (decree n° 24 of August 1997) and the total number of lots
177 increased to 537, with area of each lot between 25 and 135 ha (mean lot size = 64 ha). The
178 Matupi settlement has nine access roads (known as “ramais”): Nova Vida, Bela Vista, Matupi,
179 Matupiri, Santa Luzia, Boa Esperança, Maravilha, Triunfo and Bom Futuro (Supplementary
180 Material, Fig. S1). The total area of the Matupi settlement is 34,938 ha, based on a vector map
181 of the settlement’s boundary provided by INCRA.

182

183 2.2. *Mapping deforestation through 2016, identify actors and linking actors to deforestation*
 184 *patches*

185

186 We manually mapped cleared areas from 1994 to 2016 in the Matupi settlement by
 187 visual interpretation at 1:20,000 scale, where the appearance of areas in a satellite image
 188 displayed on a large high-definition computer screen is used to identify deforestation. Cleared
 189 areas mapped in a given year (e.g., 2000) were used as a mask for mapping cleared areas in the
 190 next year (e.g., 2001). The area of each polygon was then calculated and areas < 1 ha were
 191 excluded to reduce noise caused by small polygons, which means that the minimum map unit
 192 considered in our study was 1 ha.

193 Polygons (i.e., patches) of clearing for each year were delimited based on the visual
 194 appearance of the cleared areas, which reflects their spectral response. When boundaries
 195 between adjacent cleared areas were visible, then each area was mapped as a distinct polygon
 196 for the year in question. We used this refined approach since the clearing process could help
 197 distinguish the actions of different actors. Because the occupation process in the Matupi
 198 settlement started in 1995, we began mapping clearing using the 1994 Landsat image as a
 199 reference. The polygons of cleared areas mapped for 1994 therefore represent cumulative areas
 200 and those from 1995 to 2016 represent annual clearing. Additional information on methods used
 201 for mapping deforestation is available in the Supplementary Material.

202 Identification of the actors and their clearing (i.e., polygons of deforestation) was done
 203 based on the dataset for the Matupi settlement provided to us by the Amazonas office of INCRA
 204 in Manaus. This dataset consisted of (i) a vector map of lot boundaries (n = 537 lots), (ii)
 205 occupation survey (*Levantamento Ocupacional*) data on families in the Matupi settlement
 206 collected in October 2011 in 526 lots, and (iii) data on property diagnoses collected by INCRA
 207 in 164 lots from 2014 to 2016. Datasets (ii) and (iii) were obtained during *in loco* visits to the
 208 lots by an INCRA officer. In our analysis, we used information on the landholder and the
 209 beginning date of occupation for the lot. We also used data obtained during our fieldwork in
 210 2016, which consisted of GPS points of the lot boundaries on the six access roads we visited
 211 (Matupi, Matupiri, Maravilha, Triunfo, Bom Futuro and Nova Vida; Supplementary Material,
 212 Fig. S1).

213 All of these data assisted us in identifying and spatially locating the landholders and
 214 their polygons of deforestation. Thus, for example, if data in INCRA's 2011 occupation survey
 215 indicated that a landholder had occupied a given lot since 2004, then the polygons of
 216 deforestation from 2004 to 2011 were attributed to that landholder. In addition, if the same
 217 landholder occupied the lot in 2016, then any 2012-2016 deforestation polygons were also
 218 attributed to the landholder. When the year of occupation was not mentioned, only polygons of
 219 deforestation from 2011 were attributed to the landholder. We used this approach to be sure of
 220 correctly associating the actor and his or her clearing in the lot because the polygons of
 221 deforestation in a lot could be made by different actors who occupied the lot at different times.
 222 Out of a total of 2551 polygons of deforestation mapped in the Matupi settlement, we could
 223 identify the actors in 732 polygons (29%). We performed chi-square and Fisher's tests to assess
 224 the association between the actor type and deforestation pattern based on the samples of
 225 polygons where we identified both the actors and their clearing patterns. For classification of
 226 the deforestation pattern, we used 164 of the identified polygons as the dataset, which we
 227 divided between training and validation samples.

228

229 2.3. *Classification of deforestation patterns*

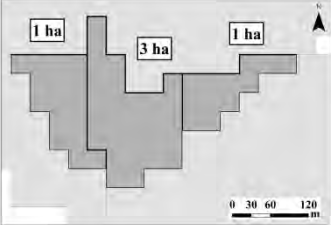
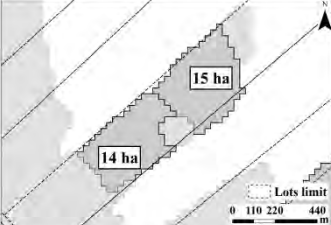
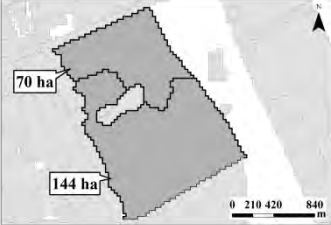
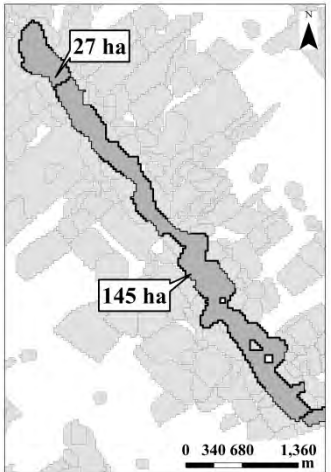
230

231 The method used to classify deforestation patterns was based on the deforestation
 232 polygons mapped in the previous step. We used the GeoDMA (Geographic Data Mining

233 Analyst, Version 0.22a) plugin (Körting et al., 2013) in Terra View 4.2.2 software to classify
234 the patterns of deforestation. The classification steps consisted of (i) feature extraction based on
235 the characteristics of deforestation patches (i.e., polygons), where patch metrics (size and shape
236 of polygons) were calculated for each patch and stored in the attribute table of the deforestation
237 vector map, (ii) selection of patch samples in which we only included patches where the actor
238 type was known based on INCRA data and where the previously defined deforestation spatial
239 pattern as defined in Table 1 was also known, (iii) classification of all deforestation patches by
240 running the C4.5 data-mining algorithm for decision-tree classification (Quinlan, 1993), and (iv)
241 assessment of the classification. The typology of deforestation patterns was determined based
242 on exploratory visual analysis by superimposing the vector map of deforestation on the vector
243 map of lot boundaries and based on the authors' previous knowledge of actor types and their
244 clearing behavior from field observations in the settlement project.

245

246
247

Deforestation pattern	Actors associated with the pattern	Description
<p><i>Small irregular</i></p> 	<p>This is the most common pattern for landholders who do not concentrate lots;</p>	<p>Main activity: cattle ranching and agriculture;</p> <p>Small patches (either grouped or isolated) indicate a small clearing each year inside of the lot. Cleared areas are for pasture or agriculture.</p>
<p><i>Small geometric</i></p> 	<p>This is most common in landholders who do not concentrate lots. The cleared areas are small and respect the boundary of the lot.</p>	<p>Main activity: cattle ranching and agriculture;</p> <p>Patches can be isolated, which could be associated with the new pasture areas or grouped with older patches that could indicate the expansion of pasture.</p>
<p><i>Large geometric</i></p> 	<p>This is a predominant pattern in landholdings of individual and family landholders who have concentrated lots.</p>	<p>Main activity: cattle ranching</p> <p>Large areas cleared in one year by actors who concentrate lots.</p>
<p><i>Large irregular</i></p> 	<p>We assumed that this pattern is mainly associated with the first families or individuals who occupied the settlement, each receiving a single lot from INCRA.</p>	<p>Main activity: cattle ranching and agriculture;</p> <p>This pattern represents the beginning of the occupation process along access roads in the Matupi settlement. The occupation is characterized by clearing at the front of the lots, which can have the effect of indicating land tenure.</p>

248
249
250

Table 1 Deforestation patterns in the Matupi settlement.

251 We separated the classification into two periods: (i) 1994 to 1999 and (ii) 2000 to 2016.
 252 This was done because the initial process of occupation in the Matupi settlement resulted in large
 253 polygons of deforestation (large irregular) that could be confused with similar polygons
 254 deforested in recent years (large geometric) (Table 1). The separation into these periods results
 255 in better distinguishing the process of deforestation and the types of actors. The large irregular
 256 areas cleared along access roads in the first years are the result of the first landholders who
 257 occupied the lots each clearing the front of the lot to indicate land tenure. We could not
 258 differentiate the clearing done by these landholders in the satellite images. In contrast, the large
 259 geometric polygons cleared in recent years are attributed to lot concentration when the polygons
 260 span several lots.

261 In total, 239 polygons were used to assist the classifications. Out of this total, in 164
 262 polygons the actors who cleared them were known, and for 61 polygons we have no information
 263 about the actors (these polygons were used only for the first classification period). In the first
 264 classification period (1994-1999) we considered the “large irregular,” “small geometric” and
 265 “small irregular” patterns. For the second classification period (2000-2016) we considered the
 266 “small irregular,” “small geometric” and “large geometric” patterns (Table 1). The “small
 267 irregular” (n = 62) and “small geometric” (n = 66) patterns were in areas with the non-
 268 concentrating actor type. The “large geometric” cases (n = 22) were in areas of lot
 269 concentration.

270

271 *2.4. Estimation of lot concentration in 2011 and deforestation rates by landholders*

272

273 Since the data from the 2011 occupation survey of families covered most of the lots in
 274 the Matupi settlement, we used these data to classify the vector map of lot limits for each actor
 275 type. The actors were divided into two major groups: non-concentrators and concentrators, the
 276 latter group including both individual and family actor types. When concentration in
 277 neighboring lots was found, we merged these lots into one representing the landholding of a
 278 concentrator. For non-concentrators, the landholding and the lot area are the same. We use the
 279 term “landholding” to refer to the area occupied by a single actor (individual or family); the area
 280 may be one or several lots and the occupation may or may not be legal.

281 The criterion used to identify concentration by families was if the members of the same
 282 family occupied neighboring lots and one of the family members resided in the neighboring lot
 283 (e.g., a parent living in his or her child’s lot). We also considered as concentration by a family
 284 the cases where both (i) lots are occupied by people with the same surname and (ii) the
 285 polygons of deforestation they made, which were identified by the period that the landholders
 286 occupied the lots, span these two or more lots. We also considered a type of concentration of
 287 non-neighboring lots. This refers to concentrators of neighboring lots who also occupied one
 288 nearby lot on the same access road. We placed these cases in a separate category as
 289 “concentrators of non-neighboring lots” with the aim of comparing the dynamics of clearing in
 290 these lots with those of non-concentrators.

291 Lots excluded from our analyses (n = 21 lots) were those with unknown actors (4 lots),
 292 lots that were not visited by an INCRA officer due to inaccessibility (10 lots) and “community”
 293 lots (7 lots). The “community” lot refers to a lot allocated by INCRA to construct infrastructure
 294 such as a school, church and space for recreational activities (e.g., a soccer field). The clearing
 295 in the community lot is therefore not associated with a specific actor. In most cases there is one
 296 community lot per access road. On one of the access roads (Boa Esperança) the community lot
 297 was occupied by a landholder, and it was included in our analysis.

298 We then performed an intersection between the vector map of lot boundaries updated to
 299 2011 and the vector map of deforestation patterns classified to estimate 1995-2011 deforestation
 300 rates per landholder (i.e., clearing per year in the area occupied by each landholder). Although

301 we are aware that deforestation in the lot could be done by different actors who occupy the lot at
 302 different times, we consider that it is important to establish the deforestation trajectories and
 303 rates of deforestation in areas where it was known whether or not the lot was occupied by
 304 concentrators in 2011. Landholders who were identified in this analysis as occupying the lot or
 305 area (in the case of concentrators of neighboring lots) in 2011 had inherited clearing done by
 306 previous landholders. To estimate the remaining forest in 2011, deforestation from 2012 to 2016
 307 was considered to have been forest in 2011, and this total was summed with the forest in 2011.
 308 Because our dataset lacked normality, a non-parametric statistical test (Mann-Whitney U) was
 309 performed. Additional information on methods is available in the Supplementary Material.

310

311 3. Results

312

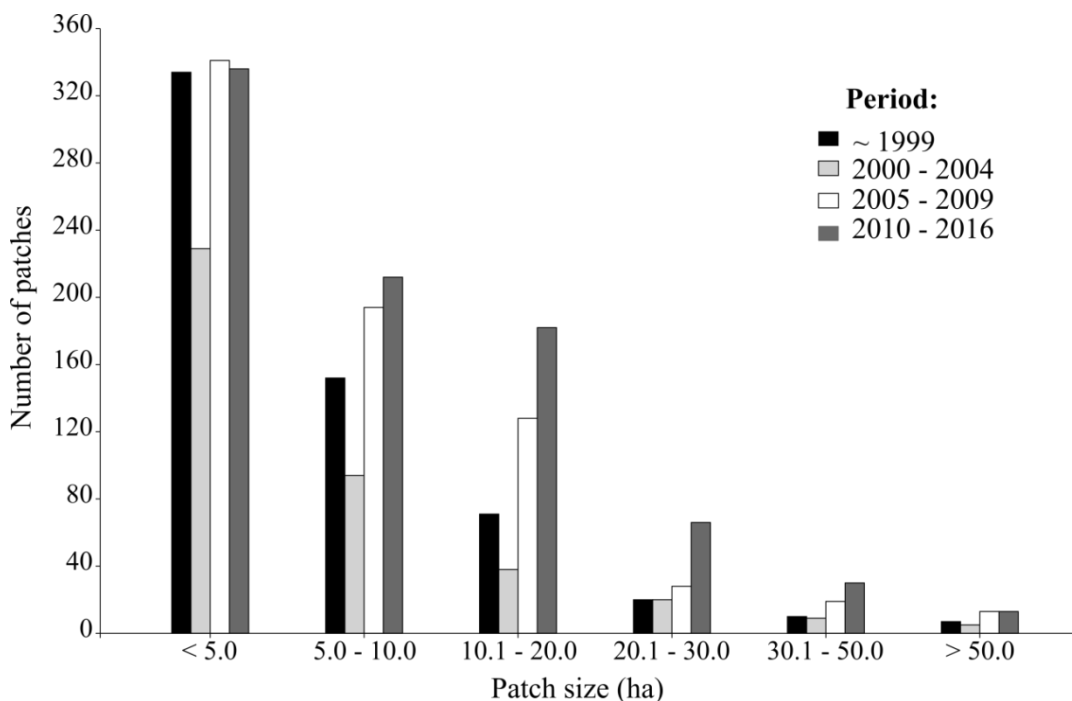
313 3.1. Spatial and temporal dynamics of deforestation

314

315 The total area cleared through 2016 in the Matupi settlement was 22,945 ha (66% of the
 316 34,938-ha settlement area), and the mean clearing per year (1995-2016) was 1026 ha. Peaks of
 317 deforestation occurred in 1997 and 2005 (9% and 10% of the total deforestation, respectively).
 318 In 2011 and 2016, high rates of deforestation were observed again, each of these years
 319 representing 8% of the total deforestation. In contrast, substantial reductions in deforestation
 320 were observed in 2006 (with a decrease of 1622 ha in relation to 2005) and in 2012 (with a
 321 decrease of 1199 ha in relation to 2011). The largest deforestation increment (1891 ha) occurred
 322 when the settlement area was officially expanded in 1997 (Supplementary Material, Fig. S1 and
 323 S2).

324 The polygons (i.e., patches) ranged from 1 ha (minimum area considered) to 167 ha. In
 325 general, as patch size increased the numbers of polygons decreased for all periods analyzed.
 326 Most patches (74% or 1892 polygons) were in the < 5 and 5 - 10 ha size ranges (Fig. 2). The
 327 2000-2004 period had the lowest number of patches in comparison with other periods for the
 328 three first classes (< 5, 5 - 10 and 10.1 - 20 ha). In contrast, the 2010-2016 period had a greater
 329 number of patches for most sizes analyzed in comparison with other periods (Fig. 2).

330



331

332

Fig. 2 Numbers of patches (polygons) of different sizes and in different periods of time.

333
 334
 335
 336
 337
 338
 339
 340
 341
 342
 343
 344
 345
 346
 347
 348
 349
 350
 351
 352
 353
 354
 355
 356
 357
 358
 359
 360
 361
 362
 363
 364

3.2. Classification of deforestation patterns by actor type

A decision tree for the first classification (1994 – 1999) identified compactness and normalized perimeter as the best landscape metrics for separating the deforestation patterns (Fig. S3). The normalized perimeter metric transformed values between the minimum and maximum perimeters into values in the interval between 0 and 1. In the second classification (2000 – 2016), compactness and area best differentiated the “small irregular” from the “small geometric” and “large geometric” patterns (Fig. S3). “Compactness” (which was used in both classifications), is a metric of patch shape (Eq. 1) that is greatest for irregular patch shapes and allows these to be separated from geometric shapes. Clearing of larger landholders can be expected to have more regular geometric patch shapes because these actors hire outside groups to clear predefined areas, rather than using family labor supplemented by individual day laborers (who may choose to avoid unfavorable topography or other obstacles). Polygons > 33.7 ha were identified as the “large geometric” pattern.

$$\text{Compactness} = (\text{perimeter}/\text{area})/\sqrt{\text{area}} \quad (1)$$

The confusion matrix in the first classification indicated that four polygons of the “small geometric” pattern were classified as “large irregular” (Supplementary Material, Table S2). The Kappa values were 0.97 (training sample versus classification) and 0.87 (validation sample versus classification). In the second classification the confusion matrix indicated that only one sample of the “large geometric” pattern was misclassified as “small geometric” and one sample of the “small geometric” pattern was misclassified as “small irregular” (Supplementary Material, Table S3). The Kappa values were 0.96 (training sample versus classification) and 1 (validation sample versus classification).

Deforestation-pattern classification through 2016 indicated that “small geometric” (44% or 9988 ha) and “small irregular” (31% or 7092 ha) were the most representative patterns in the Matupi settlement. The “large geometric” (18% or 4045 ha) and “large irregular” (8% or 1820 ha) patterns accounted for less area as of 2016 (Fig. 3).

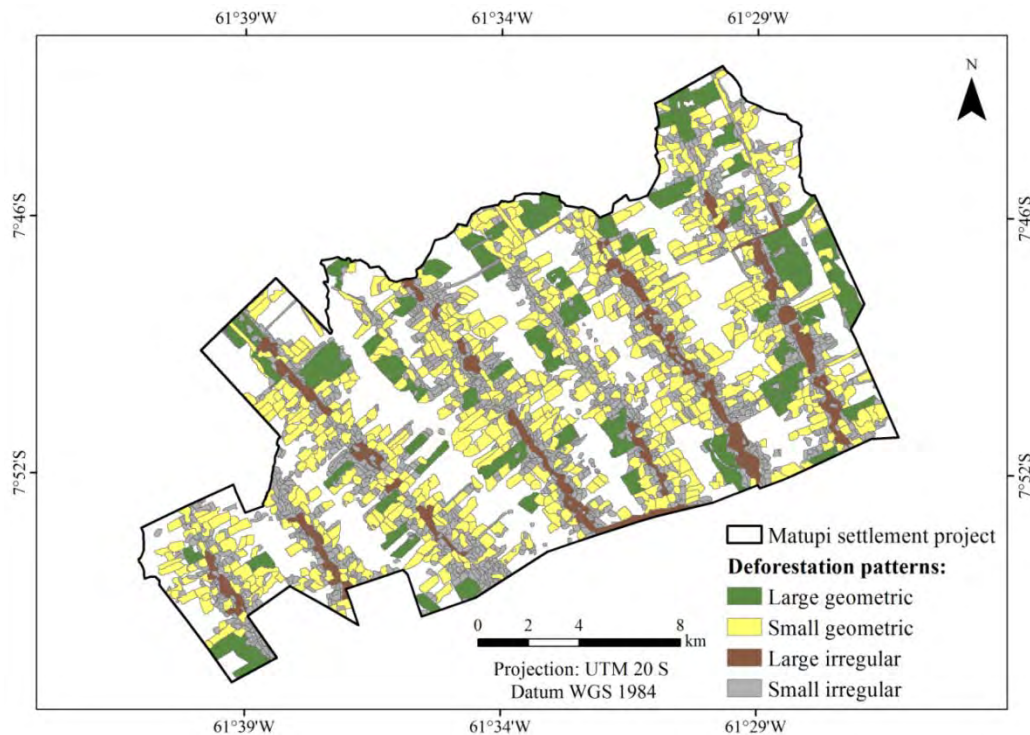
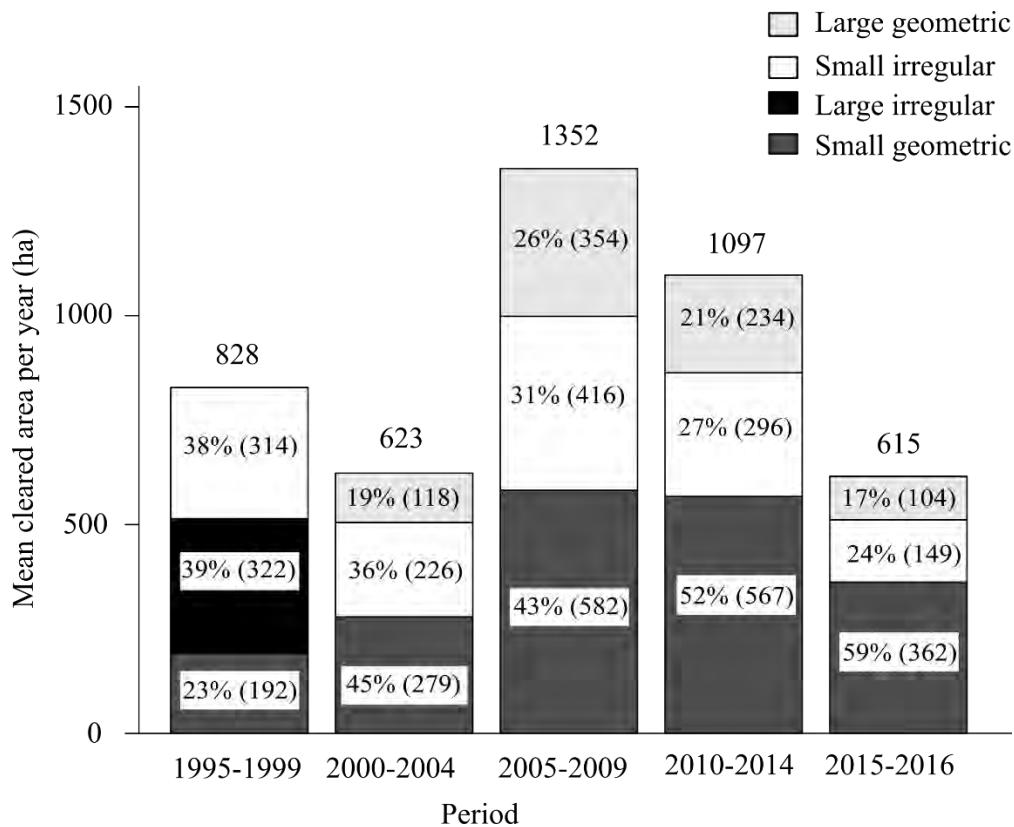


Fig. 3 Deforestation pattern classification in the Matupi settlement (2016).

“Small irregular” and “small geometric” were the patterns that encompassed the greatest numbers of patches (2428 polygons or 95% of the total). The mean size of “small irregular” polygons (4 ha) was smaller than that of the “small geometric” polygons (15 ha). However, both categories had some polygons with the same size (range = 1 – 20 ha for “small irregular” and 6 – 33.8 ha for “small geometric”). The “large irregular” pattern (mean = 34 ha) had the least polygons (54) but had the widest size range (13 – 145 ha). “Large geometric” (mean = 59 ha) also encompassed a wide range of polygon sizes (34.1 – 167 ha), with some polygons larger than those in the “large irregular” category (Fig. S4).

3.3. Temporal dynamics of deforestation patterns

The mean contribution per year of each deforestation pattern to the total for the Matupi settlement over the period from 1995 to 2016 indicated that “large irregular” was the pattern with the largest area cleared per year (322 ha) from 1995 to 1999, followed by “small irregular” with a mean of 314 ha per year (Fig. 4). Since 2000 the mean area of the “small geometric” type cleared per year was the largest in comparison with the other patterns. The mean area cleared per year in the “small geometric” pattern increased progressively from the 1995-1999 period to the 2010-2014 period, followed by a decrease in the 2015-2016 period. The “large geometric” pattern did not exist prior to 2000, so we only included this pattern from 2000 onwards (excluding it from the earlier period avoids confusion with the initial contiguous clearings along the access roads at the fronts of the lots). Since 2000 the “large geometric” pattern had an increase in the annual mean, rising from the 2000-2004 period to the 2005-2009 period and decreasing in the subsequent periods. “Small irregular” followed the same trend as “large geometric” but with greater mean areas cleared per year in all of the periods.

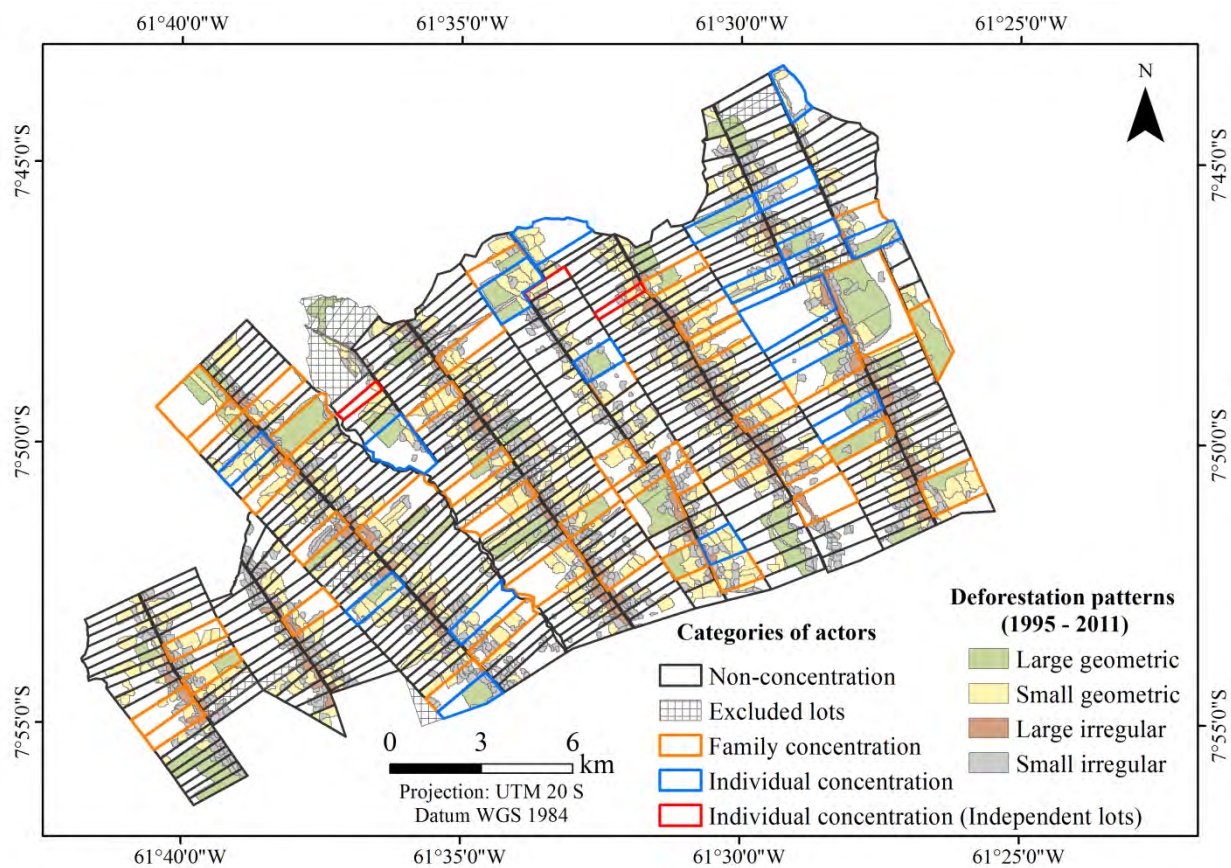


393
394 **Fig. 4** Mean area cleared per year for each time interval and deforestation pattern. Values in
395 parentheses represent the areas in hectares.

396
397 *3.4. Lot concentration (2011) and deforestation rates (1995-2011) by actor type*

398
399 Lot concentration by individuals and families was found in 152 lots or 29% of the total
400 analyzed (n = 516 lots). The area covered by landholders who concentrated lots represented
401 28% (9653 ha) of the settlement area (Fig. 5). Out of this total, 68% (6546 ha) represented
402 concentration by families (n = 42 families and 105 lots concentrated) and 32% (3107 ha) by
403 individuals (n = 18 individuals and 47 lots concentrated). The numbers of lots concentrated
404 ranged from two to ten, with the most frequent number being two lots. Of the total area
405 concentrated by actors with two lots (5905 ha), families represented 69% (4065 ha) and
406 individuals 31% (1840 ha) (Fig. 5 and Table 2).

407



408
409
410
411
412

Fig. 5 Boundaries of landholdings updated to 2011 and deforestation patterns (1995 - 2011). The “Individual concentration (Independent lots)” is the same as “concentrators of non-neighboring lots” mentioned in the text.

Concentration category	Numbers of landholders (concentrators)	Number of lots concentrated per landholder	Total numbers of lots concentrated for each actor type	Minimum and maximum areas of landholdings (ha)	Mean area of landholdings (ha)	Total area concentrated (ha)
Individual	14	2	28	118.5-163.3	131.4	1,840.2
Individual (non-neighboring lots, in addition to contiguous lots counted above)	4	4	16	229.9-398.7	276.3	1,105.3
	3	1	3	49.5 - 60.2	53.9	161.8
	33	2	66	55.7-194.4	123.2	4,064.5
	4	3	12	181.8-246.3	213.0	852.0
Family	3	4	12	232.5-272.7	246.4	739.1
	1	5	5	-	-	291.3
	1	10	10	-	-	599.0
Total			152			9,652.8

414 **Table 2.** Types of concentration found in 2011 and numbers of lots concentrated in the Matupi
 415 settlement.

416

417

418

419

420

421

422

423

424

425

426

427

428

429

430

431

432

433

434

435

436

437

438

439

440

441

442

443

Landholders with one lot were the largest category in terms of numbers (364 lots or 71%). The total area covered by this category was 23,517 ha or 68% of the Matupi settlement area in 2011 (34,796 ha, based on the vector map of lot boundaries) (Fig. 5). The sizes of the lots of non-concentrating actors ranged from 40.5 to 134.6 ha (mean = 64.6 ha).

Non-concentrators and concentrators of non-neighboring lots had similar mean annual clearing per landholding from 1995 to 2011, the annual rates being 1.7 ± 1.2 ha (mean \pm SD) and 1.2 ± 1.5 ha, respectively. Concentrators of two lots had similar mean rates per year whether the concentration was by families (4.1 ± 2.8 ha) or individuals (5.1 ± 4.6 ha). Mean annual clearing per landholding in the case of families was similar for concentrators of three lots (9.0 ± 12.8 ha) and four lots (9.6 ± 11.3 ha), but individuals with four lots had a slightly lower mean rate (7.2 ± 8.8 ha) in comparison with families with the same numbers of lots (9.6 ± 11.3 ha). A family concentrating five lots had a lower mean (6.2 ± 12.2 ha) compared to those with three or four lots, and a family with ten lots had the highest mean (23.9 ± 38.7 ha).

The mean annual clearing from 1995 to 2011 per landholding indicated significant differences in all pairwise tests ($p < 0.001$) in comparing non-concentrators ($n = 364$ landholders or lots) with concentrators of two lots ($n = 47$ concentrators) and of three or more lots ($n = 13$ concentrators) (Fig. S5). Similarly, the mean annual clearing per lot for the same period showed significant differences ($p < 0.001$) in comparing non-concentrators (1.7 ± 1.2 ha) with concentrators of two lots (2.2 ± 0.8 ha) and of three or more lots (2.2 ± 0.9 ha). No significant differences ($p = 0.54$) were found in the mean annual clearing per lot between concentrators of two and three or more lots (Table 3). However, when concentrators were analyzed separately in categories distinguishing families and individuals and the numbers of lots concentrated, we found that non-concentrators and three types of concentrators did not differ significantly ($p > 0.05$) in their mean annual clearing per lot. The categories were a family concentrator of 5 lots, individual concentrators of 4 lots and family and individual concentrators of non-neighboring lots (Supplementary Material, Tables S4 and S5).

444
445

Actor category	Total no. of lots	Mean annual clearing per lot	SD	Mean total clearing per lot
Concentrators				
Concentrators of 2 lots	94	2.2	0.8	37.3
Concentrators of 3-10 lots	55	2.2	0.9	38.1
Non- concentrators	364	1.7	0.8	29.5

446 **Table 3.** Deforestation rate per lot from 1995 to 2011 in three groups of actors categories.

447

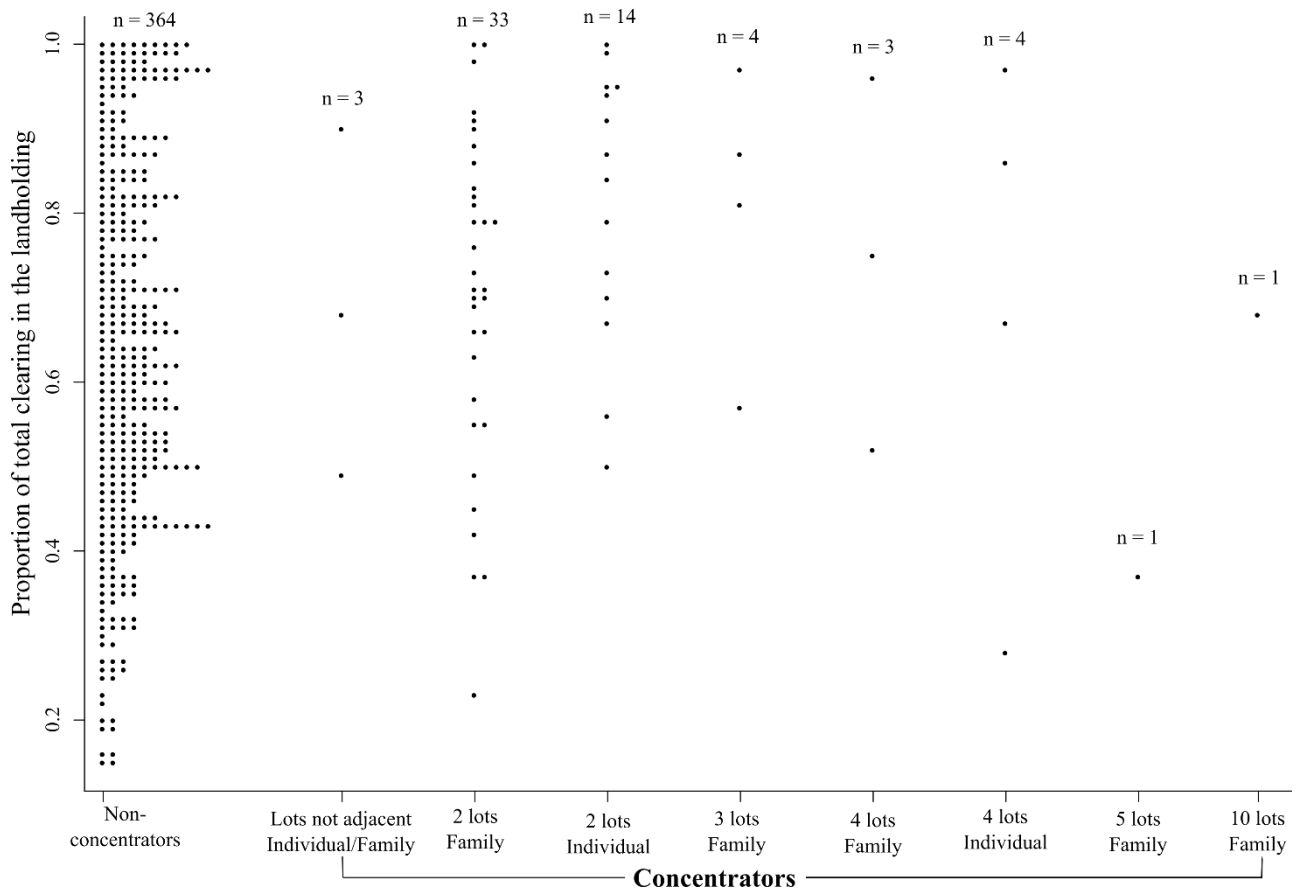
448

449 In general, non-concentrators and concentrators of non-neighboring lots had less
 450 clearing in comparison with concentrators of neighboring lots. From 1995 to 2011 the total area
 451 cleared by non-concentrators was 10,750 ha and the mean clearing per landholding was 30 ha.
 452 For concentrators of non-neighboring lots the total area cleared was 64 ha and the mean clearing
 453 per landholding was 21 ha. The total clearing (1995-2011) in the lots of non-concentrators
 454 ranged from 4 to 73 ha per lot and for concentrators of non-neighboring lots the total clearing
 455 ranged from 8 to 30 ha per lot. For concentrators (families and individuals) of two adjacent lots,
 456 the total area cleared was 3504 ha and the mean clearing per landholding was 75 ha. The total
 457 area cleared (1995-2011) per landholder of this category ranged from 21 to 128 ha. The total
 458 clearing by concentrators of three lots was 609 ha, with the mean clearing per landholding being
 459 152 ha and the total area cleared per landholding ranging from 134 to 181 ha. In the case of
 460 concentrators of four lots, the total area cleared was 978 ha with mean clearing per landholding
 461 of 140 ha and the clearing per landholding ranging from 79 to 222 ha.

461

462

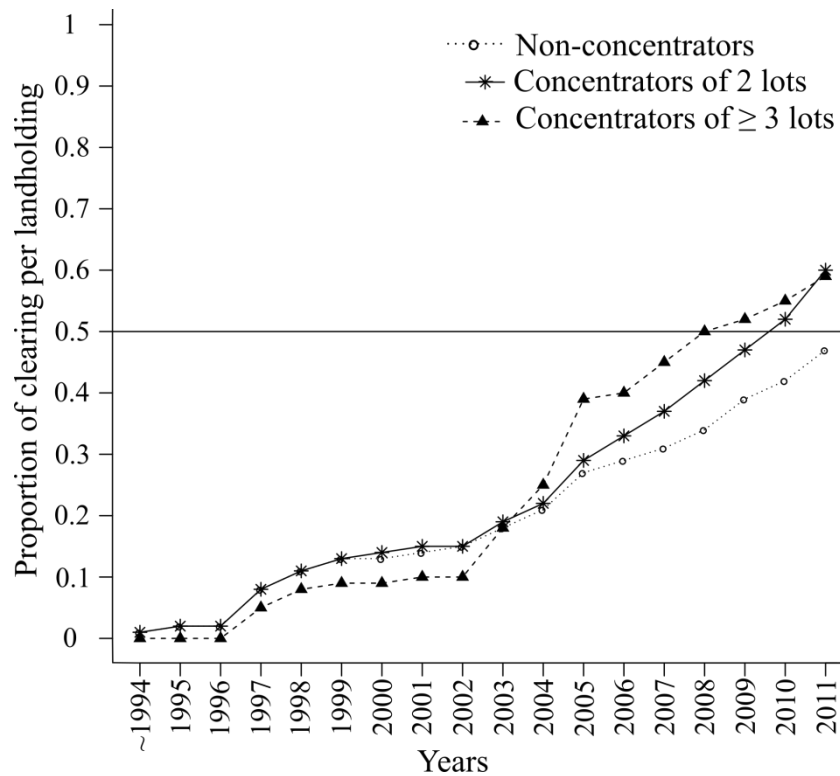
463 Only 2% of non-concentrators (n = 8 landholders) had <20% clearing in their lots (i.e.,
 464 in accordance with the Forest Code). All concentrators had total clearing >20% in the
 465 landholdings that they occupied (Fig. 6). Furthermore, 74% (n = 268) of non-concentrators had
 466 cleared more than 50% of their lots. In the landholdings of concentrators, the percentages of
 467 landholdings with more than half of their area cleared were: 87% (n = 41) for concentrators of 2
 468 lots, 100% (n = 4) for concentrators of 3 lots and 86% (n = 6) for concentrators of 4 lots. The
 family that concentrated 5 lots had less clearing (37%) in comparison with most of the
 concentrators. The family with 10 lots had 68% clearing in the landholding (Fig. 6).



469 **Fig. 6** Proportion of total area cleared per landholding from 1995 to 2011 (n = number of
 470 landholders per category).
 471
 472

473 However, because non-concentrating landholders were numerous, their contribution to
 474 total deforestation was greater (63% or 11,047 ha of the 17,426-ha total deforestation through
 475 2011), as well as per year, as compared to the total for landholders who concentrate lots (Fig.
 476 S6).

477 The proportion of area cleared through 2004 was similar for landholders with one and
 478 two lots (Fig. 7). After 2004, deforestation in areas of concentrators of two lots increased more,
 479 and in 2010 the clearing reached half of the total area of the landholdings in this category.
 480 Though 2011, areas cleared by non-concentrators still represented less than half of the total area
 481 of landholdings of this category. In areas cleared by of concentrators of ≥ 3 lots, the proportion
 482 deforested per landholding was lower through 2002 compared with other categories. However,
 483 since 2004 the proportion of clearing in this category increased and reached half of the total area
 484 of landholdings occupied by this category in 2008, which is earlier than the years for reaching
 485 this benchmark in the case of categories with fewer lots per landholder (Fig. 7).
 486



487

488

Fig. 7 Trajectory of deforestation through time per area occupied by each type of landholder.

489

Proportion of clearing represents the proportion of clearing in relation to the total area occupied by the category.

490

491

492

The remaining forest in 2011 (17,370 ha) in areas of non-concentrators represented 72% (12,471 ha) of the total forest in 2011. For concentrators, remaining forest represented 23% (3938 ha) of the total forest in the Matupi settlement in 2011. The rest of the remaining forest (5% or 961 ha) was in lots that were excluded from our analyses.

493

494

495

496

497

Considering the percentage of forest per landholding for different actor categories, we found that non-concentrators and individual concentrators of four lots had similar results for the mean percentage of forest per landholding (Table 4). Family concentrators of three and four lots had the lowest mean percentage of forest per landholding, followed by a family concentrator of ten lots and individual concentrators of two lots. In contrast, the family concentrator of five lots had the greatest percentage of forest in the landholding (Table 4). This result suggests that landholding size is not related to the proportion of remaining forest in the landholding.

500

501

502

503

504
505

Actor type (n = number of landholders analyzed)	Deforestation through 2011 (ha)	Forest in 2011 (ha)	Percentage of forest (2011) per landholding (mean \pm SD)
Non-concentrator (n = 364)	11,047 (47%)	12,471 (53%)	52.5 \pm 20.6
Concentration by individuals:			
non-neighboring lots (n = 3)	64 (40%)	98 (60%)	61.0 \pm 20.2
2 lots (n = 14)	1,213 (66%)	628 (34%)	32.8 \pm 21.0
4 lots (n = 4)	488 (44%)	617 (56%)	52.3 \pm 31.8
Concentration by families:			
2 lots (family) (n = 33)	2,332 (57%)	1,732 (43%)	41.4 \pm 20.6
3 and 4 lots (family) (n = 7)	1,108 (70%)	483 (30%)	28.3 \pm 21.0
5 lots (family) (n = 1)	105 (36%)	186 (64%)	63.9
10 lots (family) (n = 1)	405 (68%)	194 (32%)	32.4

506

507 **Table 4.** Areas and percentages of forest and deforestation per landholding of the different actor
508 types.

509

510 **4. Discussion**

511

512 *4.1. Landholding size and actor type*

513

514 Our study focused on better understanding the lot concentration process and how it
515 results in different actor types having distinct forest-clearing patterns. The mean annual clearing
516 per landholding is an important indicator of the environmental impact of accommodating the
517 different actor groups in the settlement. However, it is also important to assess deforestation rate
518 per lot since it reflects the impact of lot concentration on the overall rate of deforestation
519 (Carrero and Fearnside, 2011). We found that concentrators clear more per lot than non-
520 concentrators, which speeds deforestation.

521

522 Our finding that, in general, non-concentrators (mean lot size = 65 ha) had a higher
523 percentage of remaining forest than concentrators is similar to the observations of Godar et al.
524 (2012b), who found that actors who focused on cattle ranching with property sizes from 200 to
525 600 ha (and who were more capitalized) had less remaining forest in their properties in
526 comparison with less-capitalized colonists with property sizes under 200 ha. In addition, a
527 recent study in the Ouro Preto do Oeste settlement in Rondônia found that actors who
528 deforested more for cropland or pasture (the main income activities) obtained larger incomes
529 than those who deforested less. This is because clearing is linked to accumulation of household
530 assets (Mullan et al., 2018). Thus, the income from pasture expansion is a motivation for asset
531 accumulation that could be self-perpetuating for actors who concentrate land (Mullan et al.,
532 2018).

532

533 We did not find major differences between annual clearing per landholding of family
534 and individual concentrators. This suggests that the number of lots concentrated has more
weight in the dynamics of clearing than does the type of concentration (family versus

535 individual). In addition, the INCRA dataset reported (and we also found in the fieldwork) that
 536 cases (n= 10) of family concentration exist where a single member of the family is responsible
 537 for clearing in the landholding. The other family members either work in activities not directly
 538 related to production in the landholding or live outside of the settlement. Thus, in practice,
 539 decisions about clearing are made by one person. In our study, out of a total of 42 cases of
 540 family concentrators (105 lots: Table 3), where, in general, each member of the family occupies
 541 one lot, 44% (46 members of the concentrator families) lived in the settlement according to data
 542 in the INCRA occupation survey conducted in 2011. For individual landholders who
 543 concentrated neighboring lots, out of a total of 18 landholders (concentrating a total of 44 lots),
 544 44% (8 landholders) lived in the settlement in one of the lots they occupied.

545 The small area cleared by concentrators of 2 and ≥ 3 lots was mostly cleared between
 546 1995 and 2002. This suggests that the process of lot concentration started mainly in 2003, or
 547 eight years after the initial occupation of the settlement, and that the clearing before 2003 in the
 548 concentrated lots had been done by the previous landholders.

549 Similarly, da Silva (2012) found that 7.3 years is the average residence time of
 550 landholders in the Matupi settlement and only 3% of landholders interviewed were originally
 551 settled by INCRA. A similar trend was observed in a settlement located in Vale do Anari (in the
 552 state of Rondônia), where during the first six years of settlement occupation, cleared areas were
 553 concentrated near access roads, and patches had irregular linear patterns. After this early stage,
 554 medium and large landholders bought lots from previous settlers to establish cattle ranches.
 555 Large clearings started to appear and increased gradually through time as result of lot
 556 concentration. The patches associated with these landholders were > 50 ha in area (dos Santos
 557 Silva et al., 2008).

558 559 *4.2. Small patches of deforestation*

560
561 Our study found a total of 22,945 ha of clearing in the Matupi settlement, whereas
 562 PRODES estimated an area of 21,504 ha through 2016 (Brazil, INPE, 2018a). We mapped 1441
 563 ha (6.7%) more clearing than PRODES. This could be due the larger minimum area detected by
 564 PRODES (6.25 ha) as compared to our study (1 ha), and because we considered roads as
 565 clearing (when visible in the Landsat images). The difference could also be at least partly a
 566 result of the different image dates used as the reference for the mapping (30 July 2016 by
 567 PRODES versus 12 August 2016 in our study). In addition, because we discriminated clearing
 568 considering the spectral response of land-cover change (i.e., clearcut, initial regeneration after
 569 clearcutting, and slash-and-burn), the numbers of small polygons increased by 19% (416
 570 polygons), raising the total from 2135 (if interpretation was done without feature discrimination
 571 of clearing) to 2551 polygons. The small polygons were classified mainly as “small geometric”
 572 and “small irregular” patches. We decided to use the feature-discrimination approach because
 573 size and shape of patches are important metrics for differentiating the patterns and because this
 574 approach reduced the overestimation of area that occurs when we associate actors with
 575 polygons, in comparison to mapping without this discrimination. The result was therefore more
 576 detailed and achieved a better separation of deforestation that occurred in nearby areas in the
 577 same year but was done by different landholders.

578 A recent study has found a pervasive rise in small-scale deforestation in Brazilian
 579 Amazonia as a whole (Kalamandeen et al., 2018). Despite differences of scale between our
 580 study (local scale) and the study by Kalamandeen et al. (2018) (regional scale), we found a
 581 similar overall tendency, demonstrating that (i) as patch size increases the number of patches
 582 decreases and (ii) the contribution of small patches has increased through time.

583 In Brazil’s Legal Amazonia region, Escada et al. (2011) found that, of the 6646 km²
 584 deforested in 2009, 60% (4003 km²) was in patches <25 ha in area while only 1.7% (113 km²)

585 was in patches >1000 ha in area. The same study found that the percentage of deforestation in
 586 patches <25 ha in size increased from 22% (5897 km² out of 21,650 km² of deforestation) in
 587 2002 to the 60% found in 2009. For annual clearing in Legal Amazonia in the same period,
 588 Rosa et al. (2012) found that patches 6.25-50 ha in area increased from 30% (6495 km²) in 2002
 589 to 73% (5449 km²) in 2009. Rosa et al. (2012) suggested that the decline of large patches could
 590 be attributed to the historic trajectory of deforestation in some municipalities, lower
 591 deforestation rates being reflected in the smaller size of patches in recent deforestation. In
 592 addition, Rosa et al. (2012) suggested that some landholders changed their behavior to avoid
 593 detection by environmental monitoring, clearing small patches instead of large areas. Another
 594 factor that could contribute to the increase of small patches is fragmentation of some lots into
 595 smaller landholdings, despite the fact that the much more common pattern is one of
 596 consolidation of lots (i.e., incorporation of several lots in one landholding), as reported by
 597 D'Antona et al. (2011) in a rural settlement near Santarém (Pará). These authors found that, out
 598 of a total of 587 lots analyzed, 39 (7%) were fragmented into landholdings smaller than the
 599 original lot size, 4% were fragmented and partially merged with larger landholdings and 67% of
 600 the lots were merged in large landholdings without being fragmented. Although we lack
 601 information that would allow analysis of fragmentation of previously concentrated lots, we
 602 estimated by visual interpretation that there were 30 lots in the Matupi settlement that had been
 603 occupied by non-concentrators in 2011 (Supplementary Material, Fig. S7). This could be a
 604 result of fragmentation of previously concentrated landholdings into individual lots, which is
 605 one of the processes reported near Santarém by D'Antona et al. (2011).

606

607 *4.3. 'Peaks' and 'valleys' in observed deforestation*

608

609 We observed three important phases in the deforestation trajectory in the Matupi
 610 settlement. The first phase refers to an initial occupation process (1994 to 1996) with the arrival
 611 of the first settlers. In this phase, clearing started to appear mainly as small patches in the lots in
 612 the access roads nearest their connections to the Transamazon Highway, indicating that these
 613 lots were the first lots occupied. A study in Altamira (in Pará state) reported that landholders
 614 cleared 2 to 5 ha per year in the initial stages of settlement (McCracken et al., 1999).

615 The second phase started with the official increase of settlement area in 1997, resulting
 616 in an increase in the number of lots from 465 to 537. This represents occupation of lots by new
 617 landholders settled by INCRA. Clearing is done first at the lot front both to indicate land tenure
 618 and due the convenience of proximity to the access road. The "large irregular" pattern found in
 619 the early years of occupation along the access roads reflected the clearing done at the front of
 620 each lot. Clearing declined from 2000 to 2002, with values similar to the first phase. Only a few
 621 landholders lived in the settlement during this period, which could indicate an abandonment of
 622 lots occupied initially.

623 The last phase occurred since 2003 when clearing started to increase with peaks and
 624 lows through 2016, indicating that deforestation dynamics were more intense during this period
 625 in comparison with the first years of settlement. Since 2003, annual deforestation increased in
 626 the Matupi settlement, with a large area being cleared by concentrators, this being added to the
 627 continued contribution of non-concentrators. Part of the clearing is legal (up to 20% of each
 628 lot); however, most of the clearing is illegal. Between 2005 and 2006, command-and-control
 629 actions by the Brazilian Institute of Environment and Natural Resources (IBAMA) were intense
 630 in the settlement. Despite this, a major peak of deforestation occurred in 2005, followed by a
 631 decrease in 2006. Fines alone are not enough to stop all illegal deforestation in the settlement.
 632 Application of a fine, or the possibility of a fine, can result in some landholders forgoing
 633 clearing, as we observed during the fieldwork. We believe that command-and-control actions
 634 are more effective in the case of landholders who live in the settlement, which is a minority of

635 landholders. For example, for non-concentrators, which is the group with the largest number of
636 actors (364 landholders), only 28% (102 landholders) lived in the settlement in 2011. In the case
637 of concentrators, 44% lived in the settlement.

638 A study by Schmitt (2015) reported that, although the effect of command-and-control is
639 low and is not enough to stop all illegal deforestation in Legal Amazonia, some of the actors
640 could be influenced by IBAMA's environmental inspection program. Thus, the decline of
641 annual rates of deforestation observed between 2008 and 2013 in Brazil's Legal Amazonia
642 region could be partially attributed to the inspection program (Schmitt, 2015). Note, however,
643 that the bulk of the region-wide deforestation decline that occurred between 2004 and 2012 is
644 explained by other factors (Fearnside, 2017).

645 The main activity in the Matupi settlement is cattle ranching, although a few families
646 plant some agricultural crops in addition to their pasture. A dairy factory began operation in
647 Matupi District in 2013, and it is currently the largest dairy factory in the state of Amazonas.
648 Landholders reported that beginning in 2010 a dairy-cattle "boom" occurred in the region. This
649 could have contributed to increased deforestation in 2010-2011. During our fieldwork we found
650 many cooling platforms used to store milk at the front of the lots, indicating that dairy cattle
651 were being raised. The milk is sold to the Matupi dairy factory. Landholders reported that dairy
652 cattle are normally confined, in contrast to beef cattle. This means that dairy-cattle ranching
653 requires less pasture area; for landholders who have only one lot it is therefore better to raise
654 dairy cattle than beef cattle. However, both types of cattle need pasture, and clearing in the lots
655 would tend to increase, even if at different speeds.

656 According to INPE's TerraClass program for quantifying land cover in deforested areas,
657 in 2014 pasture was the main land use in the Matupi settlement, encompassing 82% (14,865 ha)
658 of the total area cleared through 2013 (18,087 ha) (Brazil, INPE, 2018b). This agrees with the
659 large-scale finding of Almeida et al. (2016), who found pasture to be the main land use in Legal
660 Amazonia based on TerraClass data for 2008: out of a total of 707,274 km² that had been
661 cleared through 2007, pasture encompassed 63% (447,160 km²) in 2008 and only 5% (34,927
662 km²) was in annual crop cultivation.

663 Despite the first landholders having received financing under an INCRA program to
664 produce coffee and cacao, they did not have a structured chain to market the products, a means
665 of transportation to distribute the products or technical assistance to better manage production.
666 Lack of conditions to develop agricultural activities makes cattle ranching the best choice for
667 Matupi landholders. This situation is similar to other settlements established along the
668 Transamazon Highway, where settlements were designed without considering local limitations
669 in terms of transportation of products, local markets, soil quality and other factors (Moran,
670 1981; Smith, 1982; Fearnside, 1986; Mahar, 1989; Caviglia-Harris and Harris, 2011). Amazon
671 forest soils generally have high acidity and low natural fertility, making agriculture difficult. In
672 addition, some areas also have steep topography, which contributes to most of the deforested
673 area being used for pasture.

674 It is important to note that both increases and decreases in deforestation are influenced
675 by economic factors such as commodity prices (Fearnside, 2017) and agricultural credit
676 (Assunção et al., 2015). Deforestation rates are also influenced by political factors, such as
677 election cycles (Rodrigues-Filho et al., 2015).

678 679 *4.4. Environmental implications and future studies*

680
681 Understanding the deforestation patterns of actors in a settlement project located in a
682 region of cattle-ranching expansion can contribute to developing more refined spatial models of
683 deforestation. Deforestation rates and the sizes of patches in the main deforestation patterns

684 need to be associated with the actors in spatial models in order to simulate the contributions of
 685 these actors to future deforestation under different scenarios.

686 Our findings indicate a trend to increasing percentages of concentrators, especially
 687 concentrators of three and more lots, where “large geometric” is the predominant pattern (Fig.
 688 4). This category of actor has a substantial impact in the settlement because the clearing per
 689 year by each of these actors is larger than that of other actors, since this type of actor is more
 690 capitalized in comparison to the other types. This type of concentrator has the potential to
 691 increase its contribution to deforestation in the future. The presence of lot concentrators is one
 692 of the indications that current agrarian-reform policies are weak. The purpose of the settlements
 693 is to alleviate the social problems associated with Brazil’s large population of landless farmers
 694 and, despite loopholes, the agrarian-reform program’s regulations are designed to prevent lot
 695 concentration.

696 Next steps are to compare deforestation rates and the patterns of actors in settlements
 697 with those located outside of settlements. A suggestion for future studies is to investigate other
 698 metrics that could distinguish patches oriented in the horizontal direction (i.e., lot width) in
 699 areas of concentration and in the vertical direction (i.e., from the front to back of the lot, which
 700 is typical in non-concentrator landholdings). This distinction could better differentiate
 701 landholders with one and two lots. The addition of other metrics not related to spatial patterns
 702 could be used to better differentiate non-concentrators from concentrators of non-neighboring
 703 lots. In addition, future studies could compare the deforestation patterns associated with the
 704 actors in different settlement types, such as those in the “conventional” category (e.g., the
 705 Matupi settlement) versus those in the “environmentally differentiated” category (e.g.,
 706 Sustainable Development Projects and Agro-Extractivist Settlement Projects). In the
 707 “environmentally differentiated” category, the area is sometimes divided into lots in the same
 708 way as in the “conventional” category, but the actors have different profiles.

709 Brazil’s official position is that deforestation is under control and will be slower in the
 710 future, as outlined in the country’s commitments under the 2015 Paris Agreement (Brazil,
 711 2015). However, a variety of trends in underlying forces suggests otherwise: ever greater
 712 population, investment and infrastructure development imply more rather than less deforestation
 713 (Fearnside, 2017). In addition, there are trends toward weakening environmental licensing and
 714 downgrading protected areas, among other reversals of previous achievements in this area
 715 (Fearnside, 2016, 2018a,b). Lot consolidation increases deforestation both by increasing the
 716 clearing rate in the lots that have been consolidated into larger landholdings and by the
 717 deforestation that occurs elsewhere in Amazonia by the former Matupi settlers who have sold
 718 their land to lot concentrators and moved on to more-distant frontiers. The land-tenure
 719 concentration effect documented in the present study adds one more reason suggesting that
 720 future deforestation in Brazil’s Amazonian rural settlements will be faster than it was in the
 721 past.

722

723 **5. Conclusions**

724

725 The process of land concentration in settlement areas speeds deforestation.

726 Remote sensing methods are capable of spatially identifying concentration of three or
 727 more lots, which is characterized by large geometric deforestation patterns.

728 The number of lots concentrated is more important in affecting the speed of clearing
 729 than is the question of whether the concentration is done by families or by individuals.

730 Despite the fact that lot concentrators can clear in patterns similar to non-concentrators,
 731 non-concentrators rarely clear in patterns similar to those of landholders with large numbers of
 732 lots (i.e., clearing patches >34 ha per year).

733 Due the large number of lots occupied by non-concentrators, their contribution to total
 734 clearing was greater than that of concentrators. However, our study suggests that lot
 735 concentration is increasing through time. This process threatens to increase deforestation by a
 736 few landholders. The social effect of lot concentration on the agrarian reform program is
 737 negative, since fewer families are benefitted and the social role of equity in land distribution is
 738 not achieved.

739 Because settlement projects are intended to address the social issues surrounding
 740 Brazil's large population of landless farmers, the agrarian-reform program responsible for
 741 settlements has regulations designed to limit lot concentration. The lot concentration found in
 742 the present study indicates that government authorities need to identify the actors who
 743 concentrate lots based on their deforestation patterns and monitor the land-tenure concentration
 744 in settlement projects in Brazilian Amazonia, especially in new frontier areas where the
 745 conversion of forest to pasture is intense.

747 **Acknowledgments**

748
 749 We thank the National Institute for Research in Amazonia (INPA) (PRJ15.125), the
 750 National Institute of Science and Technology for the Environmental Services of Amazonia
 751 (INCT-SERVAMB), the National Council for Scientific and Technological Development
 752 (CNPq Proc. 573810/2008-7, 610042/2009-2), the Foundation for the Support of Research of
 753 the State of Amazonas (FAPEAM Proc. 708565) and the Brazilian Research Network on
 754 Climate Change (RedeClima) for financial and logistical support. The National Institute for
 755 Space Research (INPE) provided logistical support during the time that the first author spent at
 756 that institute processing and analyzing part of the data. Ana Paula Rezende of IDESAM
 757 (Institute for the Conservation and Sustainable Development of Amazonia), provided guidance
 758 for the fieldwork, and INCRA provided the 2011 Matupi settlement dataset. This study was
 759 financed in part by the Coordenação de Aperfeiçoamento de Pessoal de Nível Superior – Brasil
 760 (CAPES) – Finance Code 001.

762 **References**

- 763
 764 Alencar, A., Pereira, C., Castro, I., Cardoso, A., Souza, L., Costa, R., Bentes, A. J., Stella, O.,
 765 Azevedo, A., Gomes, J., Novaes, R., 2016. Desmatamento nos Assentamentos da Amazônia:
 766 Histórico, Tendências e Oportunidades. IPAM, Brasília, DF, Brazil. 93 pp. Available at
 767 [http://ipam.org.br/wp-content/uploads/2016/02/Desmatamento-nos-Assentamentos-da-](http://ipam.org.br/wp-content/uploads/2016/02/Desmatamento-nos-Assentamentos-da-Amaz%C3%B4nia.pdf)
 768 [Amaz%C3%B4nia.pdf](http://ipam.org.br/wp-content/uploads/2016/02/Desmatamento-nos-Assentamentos-da-Amaz%C3%B4nia.pdf). (Last access 8 June 2018).
- 769 Almeida, C.A.D., Coutinho, A.C., Esquerdo, J.C.D.M., Adami, M., Venturieri, A., Diniz, C.G.,
 770 Dessay, N., Durieux, L., Gomes, A.R., 2016. High spatial resolution land use and land cover
 771 mapping of the Brazilian Legal Amazon in 2008 using Landsat-5/TM and MODIS data. *Acta*
 772 *Amazon.* 46, 291-302. <https://doi.org/10.1590/1809-4392201505504>
- 773 Assunção, J., Gandour, C., Rocha, R., 2015. Deforestation slowdown in the Brazilian Amazon:
 774 prices or policies? *Environ. Dev. Econ.* 20, 697-722.
 775 <https://doi.org/10.1017/S1355770X15000078>
- 776 Branford, S., Torres, M., 2017. Temer signs law that could see millions of acres lost in the
 777 Amazon. *Mongabay*, 13 July 2017. [https://news.mongabay.com/2017/07/temer-signs-law-that-](https://news.mongabay.com/2017/07/temer-signs-law-that-could-see-millions-of-acres-lost-in-the-amazon)
 778 [could-see-millions-of-acres-lost-in-the-amazon](https://news.mongabay.com/2017/07/temer-signs-law-that-could-see-millions-of-acres-lost-in-the-amazon) (Last access 8 June 2018).

- 779 Brazil, 2015. Intended Nationally Determined Contribution towards Achieving the Objective of
780 the United Nations Framework Convention on Climate Change. 10 pp.
781 <http://www4.unfccc.int/submissions/INDC/Published%20Documents/Brazil/1/BRAZIL%20iN>
782 [DC%20english%20FINAL.pdf](http://www4.unfccc.int/submissions/INDC/Published%20Documents/Brazil/1/BRAZIL%20iNDC%20english%20FINAL.pdf). (Last access 8 June 2018).
- 783 Brazil, INPE (Instituto Nacional de Pesquisas Espaciais), 2018a. Projeto PRODES -
784 Monitoramento da Floresta Amazônica Brasileira por Satélite. INPE, São José dos Campos,
785 SP., Brazil. <http://www.dpi.inpe.br/prodesdigital/> (Last access 8 June 2018).
- 786 Brazil, INPE (Instituto Nacional de Pesquisas Espaciais), 2018b. Projeto TerraClass2014. INPE,
787 São José dos Campos, SP. http://www.inpe.br/cra/projetos_pesquisas/terraclass2014.php (Last
788 access 8 June 2018).
- 789 Brazil, PR (Presidência da República), 2017. Lei No. 13.465, de 11 de julho de 2017.
790 http://www.planalto.gov.br/ccivil_03/_ato2015-2018/2017/lei/113465.htm . (Last access 8 June
791 2018).
- 792 Browder, J.O., Pedlowski, M.A., Walker, R., Wynne, R.H., Summers, P.M., Abad, A., Becerra-
793 Cordoba, N., Mil-Homens, J., 2008. Revisiting Theories of Frontier Expansion in the Brazilian
794 Amazon: A Survey of the Colonist Farming Population in Rondônia's Post-Frontier, 1992-
795 2002. *World Dev.* 36, 1469–1492. <https://doi.org/10.1016/j.worlddev.2007.08.008>
- 796 Carrero, G.C., Fearnside, P.M., 2011. Forest clearing dynamics and the expansion of
797 landholdings in Apuí, a deforestation hotspot on Brazil's Transamazon Highway. *Ecol. Soc.*
798 16(2), art. 26. <http://www.ecologyandsociety.org/vol16/iss2/art26/> (Last access 8 June 2018).
- 799 Caviglia-Harris J.L., Harris D., 2011. The impact of settlement design on tropical deforestation
800 rates and resulting land cover patterns. *Agric. Resour. Econ. Rev.* 40, 451-470.
801 <https://doi.org/10.1017/S1068280500002896>
- 802 Caviglia-Harris, J.L., Sills, E.O., Mullan, K., 2013. Migration and mobility on the Amazon
803 frontier. *Popul. Environ.* 34, 338-369. <https://doi.org/10.1007/s11111-012-0169-1>
- 804 D'Antona, A., VanWey, L., Ludewigs, T., 2011. Polarização da estrutura fundiária e mudanças
805 no uso e na cobertura da terra na Amazônia. *Acta Amazon.* 41(2), 223-232.
806 <https://doi.org/10.1590/S0044-59672011000200006>
- 807 da Silva, V.V., Vettorazzi, C.A., Padovani, C.R., 2011. Assentamento rural e a dinâmica da
808 paisagem. In: XV Simpósio Brasileiro de Sensoriamento Remoto. Curitiba. Anais do XV
809 Simpósio Brasileiro de Sensoriamento Remoto – SBSR 2011. Instituto Nacional de Pesquisas
810 Espaciais (INPE), São José dos Campos, SP, Brazil. pp. 7000-7006.
811 <http://marte.sid.inpe.br/col/dpi.inpe.br/marte/2011/07.28.19.56/doc/p0539.pdf> (Last access 8
812 June 2018).
- 813 da Silva, V.V., 2012. Impacto das atividades produtivas na dinâmica da paisagem no
814 assentamento Matupi, estado do Amazonas. Doctoral thesis. Universidade de São Paulo,
815 Piracicaba, SP, Brazil. 117 pp. <http://www.teses.usp.br/teses/disponiveis/91/91131/tde->
816 [26102012-172251/pt-br.php](http://www.teses.usp.br/teses/disponiveis/91/91131/tde-26102012-172251/pt-br.php) . (Last access 8 June 2018).
- 817 dos Santos Silva, M.P., Câmara, G., Escada, M.I.S., Souza, R.C.M., 2008. Remote-sensing
818 image mining: Detecting agents of land-use change in tropical forest areas. *Int. J. Remote Sens.*
819 29, 4803-4822. <https://doi.org/10.1080/01431160801950634>

- 820 Escada, M.I.S., Maurano, L.S., Rennó, C.D., Amaral, S., Valeriano, D.M., 2011. Avaliação de
821 dados dos sistemas de alerta da Amazônia: DETER e SAD. In: XV Simpósio Brasileiro de
822 Sensoriamento Remoto. Curitiba. Anais do XV Simpósio Brasileiro de Sensoriamento Remoto
823 – SBSR 2011. Instituto Nacional de Pesquisas Espaciais (INPE), São José dos Campos, SP,
824 Brazil. pp. 2934-2943.
825 <http://mart.sid.inpe.br/col/dpi.inpe.br/marte/2011/06.27.16.55/doc/p1246.pdf> . (Last access 8
826 June 2018).
- 827 Fearnside, P.M., 1986. Human Carrying Capacity of the Brazilian Rainforest. Columbia
828 University Press, New York, NY. 293 pp.
- 829 Fearnside, P.M., 1987. Derrubada da floresta e roçagem de crescimento secundário em projetos
830 de colonização na Amazônia brasileira e a sua relação à capacidade de suporte humano. *Acta*
831 *Amazon.* 17(4) (suplemento), 123-141. <https://doi.org/10.1590/1809-43921987175141>
- 832 Fearnside, P.M., 1997. Environmental services as a strategy for sustainable development in
833 rural Amazonia. *Ecol. Econ.* 20, 53-70. [https://doi.org/10.1016/S0921-8009\(96\)00066-3](https://doi.org/10.1016/S0921-8009(96)00066-3)
- 834 Fearnside, P.M., 2001. Land-tenure issues as factors in environmental destruction in Brazilian
835 Amazonia: The case of southern Pará. *World Dev.* 29, 1361-1372.
836 [https://doi.org/10.1016/S0305-750X\(01\)00039-0](https://doi.org/10.1016/S0305-750X(01)00039-0)
- 837 Fearnside, P.M., 2008a. Amazon forest maintenance as a source of environmental services.
838 *Anais da Academia Brasileira de Ciências* 80, 101-114. <https://doi.org/10.1590/S0001-37652008000100006>
- 840 Fearnside P.M., 2008b. The roles and movements of actors in the deforestation of Brazilian
841 Amazonia. *Ecol. Soc.* 13, art. 23. <http://www.ecologyandsociety.org/vol13/iss1/art23/>. (Last
842 access 8 June 2018).
- 843 Fearnside, P.M., 2016. Brazilian politics threaten environmental policies. *Science* 353, 746-748.
844 <https://doi.org/10.1126/science.aag0254>
- 845 Fearnside, P.M., 2017. Deforestation of the Brazilian Amazon. In: H. Shugart (ed.) *Oxford*
846 *Research Encyclopedia of Environmental Science*. Oxford University Press, New York, NY.
847 <https://doi.org/10.1093/acrefore/9780199389414.013.102>
- 848 Fearnside, P.M., 2018a. Challenges for sustainable development in Brazilian Amazonia.
849 *Sustain. Dev.* 26, 141-149. <https://doi.org/10.1002/sd.1725>
- 850 Fearnside, P.M., 2018b. Why Brazil's new president poses an unprecedented threat to the
851 Amazon. *Yale Environment* 360, 8 November 2018. [https://e360.yale.edu/features/why-brazils-
852 new-president-poses-an-unprecedented-threat-to-the-amazon](https://e360.yale.edu/features/why-brazils-new-president-poses-an-unprecedented-threat-to-the-amazon)
- 853 Fujisaka, S., Bell, W., Thomas, N., Hurtado, L., Crawford, E., 1996. Slash-and-burn agriculture,
854 conversion to pasture, and deforestation in two Brazilian Amazon colonies. *Agric. Ecosyst.*
855 *Environ.* 59, 115-130. [https://doi.org/10.1016/0167-8809\(96\)01015-8](https://doi.org/10.1016/0167-8809(96)01015-8)
- 856 Geist, H.J., Lambin, E.F., 2001. What drives tropical deforestation? A meta-analysis of
857 proximate and underlying causes of deforestation based on subnational case study evidence.
858 *LUCC Report Series*, 4. pp. 66-72. <http://www.pik-potsdam.de/~luedeke/lucc4.pdf> . (Last
859 access 8 June 2018).

- 860 Godar, J., Tizado, E.J., Pokorny, B., Johnson, J., 2012a. Who is responsible for deforestation in
861 the Amazon? A spatially explicit analysis along the Transamazon Highway in Brazil. *For. Ecol.*
862 *Manage.* 267, 58-73. <https://doi.org/10.1016/j.foreco.2011.11.046>
- 863 Godar, J., Tizado, E.J., Pokorny, B., Johnson, J., 2012b. Typology and characterization of
864 Amazon colonists: A case study along the Transamazon highway. *Hum. Ecol.* 40, 251-267.
865 <https://doi.org/10.1007/s10745-012-9457-8>
- 866 Kalamandeen, M., Gloor, E., Mitchard, E., Quincey, D., Ziv, G., Spracklen, D., Spracklen, B.,
867 Adami, M., Aragão, L.E.O.C., Galbraith, D., 2018. Pervasive rise of small-scale deforestation in
868 Amazonia. *Sci. Rep.* 8(1), art. 1600. <https://doi.org/10.1038/s41598-018-19358-2>
- 869 Körting, T.S., Fonseca, L.M.G., Câmara, G., 2013. GeoDMA—Geographic data mining analyst.
870 *Comput. Geosci.* 57, 133-145. <https://doi.org/10.1016/j.cageo.2013.02.007>
- 871 Lu, D., Li, G., Moran, E., Hetrick, S., 2013. Spatiotemporal analysis of land-use and land-cover
872 change in the Brazilian Amazon. *Int. J. Remote Sens.* 34(16), 5953-5978.
873 <https://doi.org/10.1080/01431161.2013.802825>
- 874 Mahar, D.J., 1989. Government Policies and Deforestation in Brazil's Amazon Region. The
875 World Bank, Washington, DC. 66 pp.
876 [http://siteresources.worldbank.org/BRAZILINPOREXTN/Resources/3817166-](http://siteresources.worldbank.org/BRAZILINPOREXTN/Resources/3817166-1185895645304/4044168-1185895685298/020pub_br14.pdf)
877 [1185895645304/4044168-1185895685298/020pub_br14.pdf](http://siteresources.worldbank.org/BRAZILINPOREXTN/Resources/3817166-1185895645304/4044168-1185895685298/020pub_br14.pdf) (Last access 8 June 2018).
- 878 Martins, P.F.S., Pereira, T.Z., 2012. Cattle-raising and public credit in rural settlements in
879 Eastern Amazon. *Ecol. Indic.* 20, 316–323. <https://doi.org/10.1016/j.ecolind.2012.02.031>
- 880 McCracken, S.D., Brondizio, E.S., Nelson, D., Moran, E.F., Siqueira, A.D., Rodriguez-Pedraza,
881 C., 1999. Remote sensing and GIS at farm property level: Demography and deforestation in the
882 Brazilian Amazon. *Photogramm. Eng. Remote Sens.* 65, 1311-1320.
883 https://msu.edu/~moranef/documents/99-03_RemoteSensing_GISatFarm.pdf . (Last access 8
884 June 2018).
- 885 Moran, E.F., 1981. Developing the Amazon: The Social and Ecological Consequences of
886 Government-Directed Colonization along Brazil's Transamazon Highway. Indiana University
887 Press, Bloomington, IN. 292 pp.
- 888 Mullan, K., Sills, E., Pattanayak, S.K., Caviglia-Harris, J., 2018. Converting forests to farms:
889 the economic benefits of clearing forests in agricultural settlements in the Amazon. *Environ and*
890 *Resource Econ.* 71(2), 427-455. <https://doi.org/10.1007/s10640-017-0164-1>
- 891 Nogueira, E.M., Yanai, A.M., Fonseca, F.O., Fearnside, P.M., 2015. Carbon stock loss from
892 deforestation through 2013 in Brazilian Amazonia. *Glob. Chang. Biol.* 21, 1271-1292.
893 <https://doi.org/10.1111/gcb.12798>
- 894 Quinlan, J., 1993. C4.5: Programs for Machine Learning. Morgan Kaufmann, San Mateo, CA.
895 Available at:
896 [https://books.google.com.br/books?id=HExncpjbYroC&printsec=frontcover&hl=pt-](https://books.google.com.br/books?id=HExncpjbYroC&printsec=frontcover&hl=pt-BR&source=gbs_ge_summary_r&cad=0#v=onepage&q&f=false)
897 [BR&source=gbs_ge_summary_r&cad=0#v=onepage&q&f=false](https://books.google.com.br/books?id=HExncpjbYroC&printsec=frontcover&hl=pt-BR&source=gbs_ge_summary_r&cad=0#v=onepage&q&f=false)
- 898 Rodrigues-Filho, S., Verburg, R., Bursztyn, M., Lindoso, D., Debortoli, N., Vilhena, A.M.G.,
899 2015. Election-driven weakening of deforestation control in the Brazilian Amazon. *Land Use*
900 *Policy* 43, 111–118. <https://doi.org/10.1016/j.landusepol.2014.11.002>

- 901 Rosa, I., Souza, C., Ewers, R.M., 2012. Changes in size of deforested patches in the Brazilian
902 Amazon. *Conserv. Biol.* 26, 932-937. <https://doi.org/10.1111/j.1523-1739.2012.01901.x>
- 903 Schmitt, J., 2015. Crime sem castigo: a efetividade da fiscalização ambiental para o controle do
904 desmatamento ilegal na Amazônia. Doctoral thesis. Universidade de Brasília, Brasília, DF,
905 Brazil. 188 pp. http://repositorio.unb.br/bitstream/10482/19914/1/2015_JairSchmitt.pdf . (Last
906 access 8 June 2018).
- 907 Smith, N.J.H., 1982. *Rainforest Corridors: The Transamazon Colonization Scheme*. University
908 of California Press, Berkeley, CA. 248 pp.
- 909 Yanai, A.M., Nogueira, E.M., Graça, P.M.L.A., Fearnside, P.M., 2017. Deforestation and
910 carbon stock loss in Brazil's Amazonian settlements. *Environ. Manage.* 59, 393-409.
911 <https://doi.org/10.1007/s00267-016-0783-2>
- 912 Zipperer, W.C., 1993. Deforestation patterns and their effects on forest patches. *Landsc. Ecol.* 8,
913 177-184. <https://doi.org/10.1007/BF0012534>

SUPPLEMENTARY MATERIAL

Title: Deforestation dynamics in Brazil's Amazonian settlements: Effects of land-tenure concentration

Journal of Environmental Management <https://doi.org/10.1016/j.jenvman.2020.110555>

Authors: Aurora Miho Yanai¹, Paulo Maurício Lima de Alencastro Graça^{1,4}, Maria Isabel Sobral Escada², Leonardo Guimarães Ziccardi^{1,3}, Philip Martin Fearnside^{1,4,5}

¹Department of Environmental Dynamics, National Institute for Research in Amazonia (INPA), Av. André Araújo n° 2936, CEP 69067-375, Manaus, Amazonas, Brazil.

²Department of Image Processing, *National Institute for Space Research (INPE)*, Av. Astronautas, n° 1758, CEP 12227-010, São José dos Campos, São Paulo, Brazil

³Department of Forestry, Michigan State University, East Lansing, MI, 48824, USA

⁴Brazilian Research Network on Climate Change (RedeClima)

⁵Corresponding author

Email addresses: yanai@inpa.gov.br; pmlag@inpa.gov.br; isabel@dpi.inpe.br; leonardo.g.ziccardi@gmail.com; pmfearn@inpa.gov.br
Tel: +55(92) 3643-1822

Additional information on Materials and Methods	p. 2
Table S1. Actor-polygons found associated with classified deforestation patterns.....	p. 4
Table S2. Confusion matrix for classification from 1994 to 1999.	p. 4
Table S3. Confusion matrix for classification from 2000 to 2016.	p. 5
Table S4. Deforestation rate per lot from 1995 to 2011 for each actor category.....	p. 6
Table S5. P-values in pairwise tests comparing actor categories	p. 7
Fig. S1. Distribution of deforestation per year (1994-2016) mapped in the Matupi settlement	p. 8
Fig. S2. Total area cleared per year mapped in the Matupi settlement.....	p. 8
Fig. S3. Results of decision-tree classifications for the first (1994-1999) and second (2000-2016) classification periods.....	p. 9
Fig. S4. Distribution of patch areas for each deforestation pattern.....	p. 9
Fig. S5 Distribution of mean clearing per landholding from 1995 to 2011.....	p. 10
Fig. S6 Total area cleared in the period from 1995 to 2011 by type of actor (n = 17 years).p.	11
Fig. S7. Lots held by non-concentrators that had portions of large geometric deforestation patches spanning more than one lot in 2011	p. 11
References	p. 11

Additional information on Materials and Methods

Mapping deforestation through 2016 in the Matupi settlement

Cleared areas were mapped using Landsat-5 TM (1994 to 2011), ResourceSat-1 LISS-3 (2012) and Landsat-8 OLI images (2013 to 2016) (path: 231; row: 65). We used images from the U.S. Geological Survey (USGS), and for each year we chose the image with the least cloud cover. Least cloud cover was determined visually by satellite images preview in the Earth Explorer platform (<https://earthexplorer.usgs.gov/>). The best images were obtained during the dry season (end of May to October) in our study area. We performed an atmospheric correction using the FLAASH (Fast Line of sight Atmospheric Analysis of Hypercubes) tool available in Envi software to better differentiate land-cover change and to compare cleared areas in different years when necessary. The color composition was shortwave infrared (Red), near infrared (Green), and red (Blue).

We only mapped areas cleared by clearcut and areas of forest loss with severe fire where the spectral response was that of clearing. Areas degraded by logging or by non-severe fire were not mapped. All logging is selective in Amazonia because only large individuals of valuable species are harvested, leaving the remaining trees in the diverse rainforest standing, unlike logging in temperate and boreal areas where forests are clearcut for timber. Likewise, forest fires in Amazonia burn through the understory killing some trees, but do not result in crown fires that kill entire stands as in coniferous forests. Logged and burned areas are therefore not easily distinguished from undisturbed forest on satellite imagery, although techniques exist to identify heavily disturbed areas (e.g., Walker et al., 2020).

Data from PRODES (Project for Monitoring Amazonian Deforestation) were used to assist our mapping when doubts arose concerning specific areas and to verify the agreement between our mapping and the PRODES dataset as a whole (Brazil, INPE, 2018). PRODES is the Brazilian government's program of annual deforestation monitoring carried out by the National Institute for Space Research (INPE). We did not use the PRODES vector map because PRODES does not have annual deforestation mapping before 2000 for the Matupi settlement area and because the deforestation dataset from 2008 to 2014 had been modified with a spatial adjustment of the vector mask (i.e., cumulative deforestation from previous years) (Brazil, INPE, 2015, 2019). This spatial adjustment makes it difficult to use PRODES data for our spatial-temporal analysis in the Matupi settlement.

Training step for automatic classification of deforestation patterns

In the classification's training step 60% of the samples were randomly selected, and these samples were used to automatically create a decision-tree classification and in the validation step (40% of the samples). Assessment of the classification was done using a confusion matrix and the Kappa statistic (Körting et al., 2013). Decision-tree classification is a non-parametric supervised learning method that is relatively simple, explicit, flexible, robust with respect to nonlinear and noisy relation between input features and class labels, and that handles both discrete and continuous attributes and incomplete training data with missing values (Friedl and Brodley, 1997). The decision tree uses the C4.5 algorithm for classification rules (Quinlan, 1993). A dataset is classified based on the smaller subdivisions according to the decision framework defined by the tree, and the label of each class is added according to the leaf node (terminal node) into which the sample falls (Friedl and Brodley, 1997). Smaller decision trees are better because they are easier to understand and because the predictive accuracy tends to be higher than for large trees (Quinlan, 1996). Thus, the deforestation patterns

of patches were classified by analyzing a set of instances (i.e., a set of training samples) where the patterns were known. The decision tree then classified all of the patches in the deforestation map by learning based on the training set.

Classification of deforestation patterns

Fifteen landscape metrics were calculated by GeoDMA. These were examined both in raw form and after being normalized using the minimum and maximum values.

In the second classification period (2000-2016), we could not specify the type of actor exactly in several cases involving large polygons because polygons with the “large irregular” pattern (n = 37 polygons) covered large areas and we do not have information about all actors covered by this pattern. In two polygons we found that parts of the polygons belonged to non-concentrating landholders. In the “small geometric” category (n = 29), 3 samples were from non-concentrating landholders, and for the remaining 26 samples we do not have information about the type of actor, but the sizes and shapes of the clearings are similar to the others in the dataset. All cases of the “small irregular” pattern (n = 23), these clearings were in areas where landholders do not concentrate lots.

After the first classification (1994-1999) we had to manually reclassify 7 polygons from “large irregular” to “small geometric,” where 4 polygons were used as samples in the classification. In addition, 1 polygon classified as “large irregular” was manually reclassified to “small irregular.” We performed the reclassification because these polygons did not reflect the “large irregular” pattern (i.e., large polygons that covered more than one lot and that were located along access roads). The actor type is unknown for these reclassified polygons (and these polygons therefore were not used in the analyses that included actor types).

For the second classification period (2000-2016), concentrators of two lots had 12 polygons sampled and concentrators of three to ten lots had 10 polygons sampled. We had fewer samples of the “large geometric” pattern because the number of polygons available for use as samples was lower in comparison to the “small irregular” and “small geometric” patterns.

Actor types associated with classified deforestation polygons

Association of actor-polygons with deforestation patterns indicated that the “small geometric” and “small irregular” types were the typical patterns of non-concentrating landholders. Only 1% of the area identified as occupied by non-concentrators was classified as “large geometric” because the areas of the one polygon of this type were larger than the threshold (34 ha) that separated “small geometric” from “large geometric” (Table S1).

Although the samples that were used for classification of concentrators of two or more lots were characterized by the “large geometric” pattern, “small” patterns (geometric and irregular) were also found in these landholder types. Thus, of the 2,208-ha total area found to be held by concentrators of two lots, 51% was classified as “small geometric” and 23% as “small irregular.” For concentrators of 3 to 10 lots, “large geometric” was the predominant pattern in terms of area (66%), despite the fact that, in terms of the number of polygons, the most frequent types were “small geometric” (n = 16 polygons) and “small irregular” (n = 19; Table S1). Both the Chi-square and Fisher’s tests showed a highly significant association between deforestation patterns and actor types ($p < 0.001$).

Table S1. Actor-polygons found associated with classified deforestation patterns.

Actor category	Deforestation pattern			Total
	Area in hectares (number of polygons)			
	Large geometric	Small geometric	Small irregular	
Non-concentrators	41 (1)	1,780 (118)	1,487 (370)	3,308 (489)
Concentrators of 2 lots	581 (11)	1,122 (74)	505 (111)	2,208 (196)
Concentrators of 3 to 10 lots	707 (10)	295 (16)	67 (19)	1,069 (45)
Total	1,329 (22)	3,197 (208)	2,058 (500)	6,585 (730)

Estimation of lot concentration in 2011 and deforestation rates by landholders

In three cases of concentration (two cases encompassing two lots on the Maravilha access road and one encompassing three lots on Bom Futuro access road) we lacked information on the period that the family members occupied the lots; in these cases, we considered the group of lots to be concentrated based on the spatial distribution of deforestation polygons through 2011 in the landholding as a whole.

Generally there is one community lot per access road, but on the Triunfo and Matupiri access roads we identified two community lots in each access road, while in the Maravilha, Bom Futuro and Santa Luzia access roads there are no community lots.

Table S2. Confusion matrix for classification from 1994 to 1999. Values refer to numbers of samples (polygons) in training step and in the validation step. Total number of samples shown in **bold**.

	Pattern	Large irregular	Small geometric	Small irregular	Total	Error of omission
Reference	Large irregular	22; 15 (37)	-	-	37	-
	Small geometric	1; 3 (4)	16; 9 (25)	-	29	13.8%
	Small irregular	-	-	14; 9 (23)	23	-
	Total	41	25	23	89	
	<i>Error of commission</i>	9.8%	-	-		

Table S3. Confusion matrix for classification from 2000 to 2016. Values refer to numbers of samples in the training step and the validation step. Total number of samples shown in **bold**.

	Pattern	Large geometric	Small geometric	Small irregular	Total	<i>Error of omission</i>
Reference	Large geometric	12; 9 (21)	1; 0 (1)	-	22	4.6%
	Small geometric	-	39; 26 (65)	1; 0 (1)	66	1.5%
	Small irregular	-	-	37; 25 (62)	62	-
	Total	21	66	63	150	
	<i>Error of commission</i>	-	1.5%	1.6%		

Table S4. Deforestation rate per lot from 1995 to 2011 for each actor category.

Actor category	Total no. of lots	Mean annual rate per lot	SD	Mean total deforestation per lot (1995-2011)
Family concentrators of 2 lots	66	2.0	0.8	34.7
Individual concentrator of 2 lots	28	2.5	0.8	43.3
Family concentrators of 3 lots	12	3.0	0.5	50.7
Family concentrators of 4 lots	12	2.4	0.7	40.9
Individual concentrators of 4 lots	16	1.8	1.0	30.5
Family concentrator of 5 lots	5	1.2	0.2	21.0
Family concentrator of 10 lots	10	2.4	0.6	40.5
Individual and family concentrators of non-neighboring lots	3	1.2	0.7	21.2
Non-concentrators	364	1.7	0.8	29.5

Table S5. P-values in pairwise tests comparing actor categories. Values in **bold** indicate significant differences ($p < 0.05$). FC: Family concentrator; FCs: Family concentrators and ICs: Individual concentrators.

	FC of 10 lots	FCs of 2 lots	FCs of 3 lots	FCs of 4 lots	FC of 5 lots	FC and ICs of non- neighbor- ing lots	ICs of 2 lots	ICs of 4 lots
FCs of 2 lots	0.0950							
FCs of 3 lots	0.0290	0.0003						
FCs of 4 lots	1.0000	0.1139	0.0373					
FC of 5 lots	0.0166	0.0237	0.0018	0.0060				
FC and ICs of non- neighboring lots	0.0341	0.1488	0.0113	0.0424	0.5486			
ICs of 2 lots	1.0000	0.0109	0.0786	0.7339	0.0007	0.0210		
ICs of 4 lots	0.1873	0.2785	0.0037	0.0898	0.3623	0.5755	0.00624	
Non- concentrators	0.0030	0.0044	0.0000	0.0039	0.0749	0.2861	0.0000	0.7907

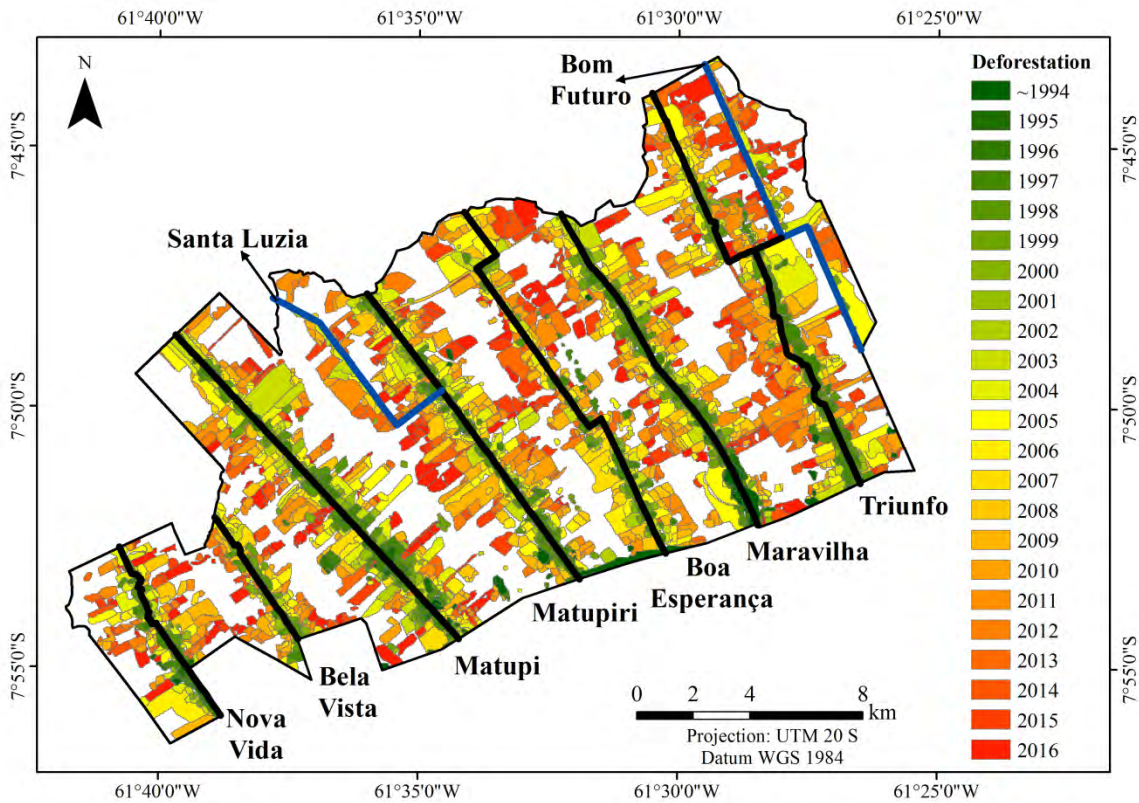


Fig. S1. Lots and access roads in the Matupi settlement, showing deforested areas and dates of deforestation.

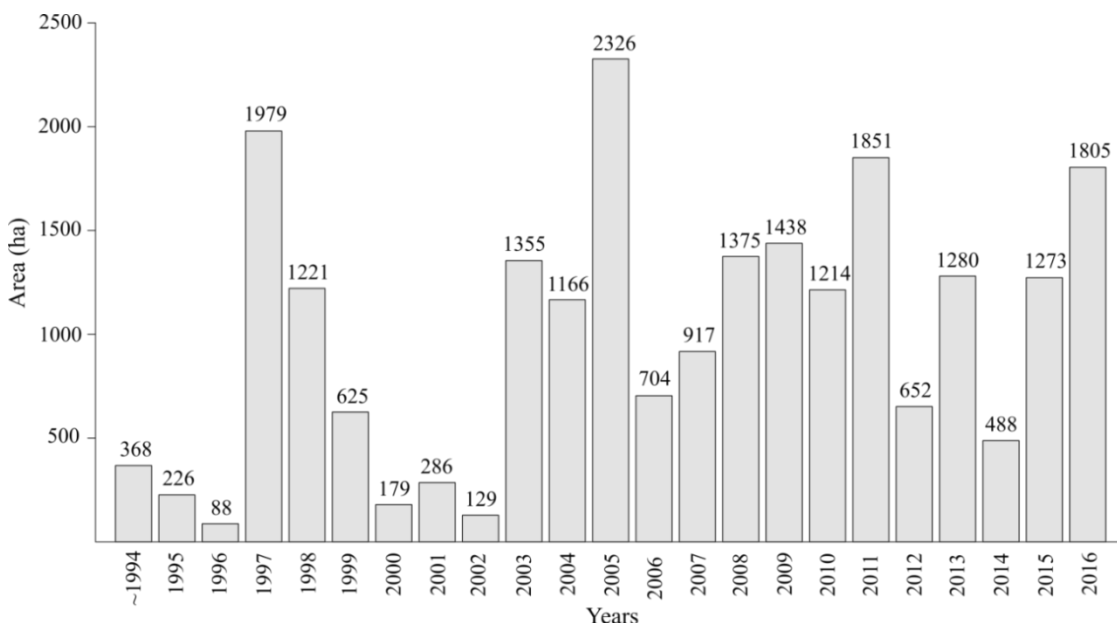


Fig. S2. Total area cleared per year mapped in the Matupi settlement.

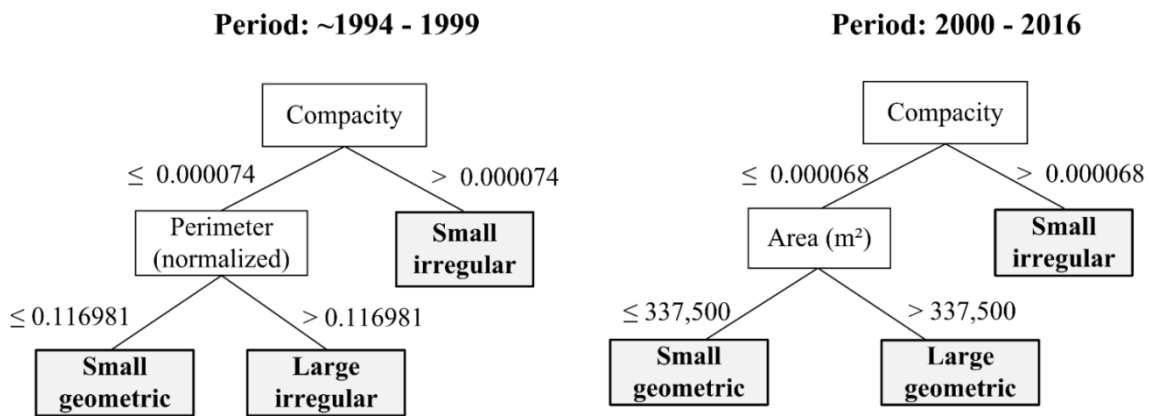


Fig. S3. Results of decision-tree classifications for the first (1994-1999) and second (2000-2016) classification periods.

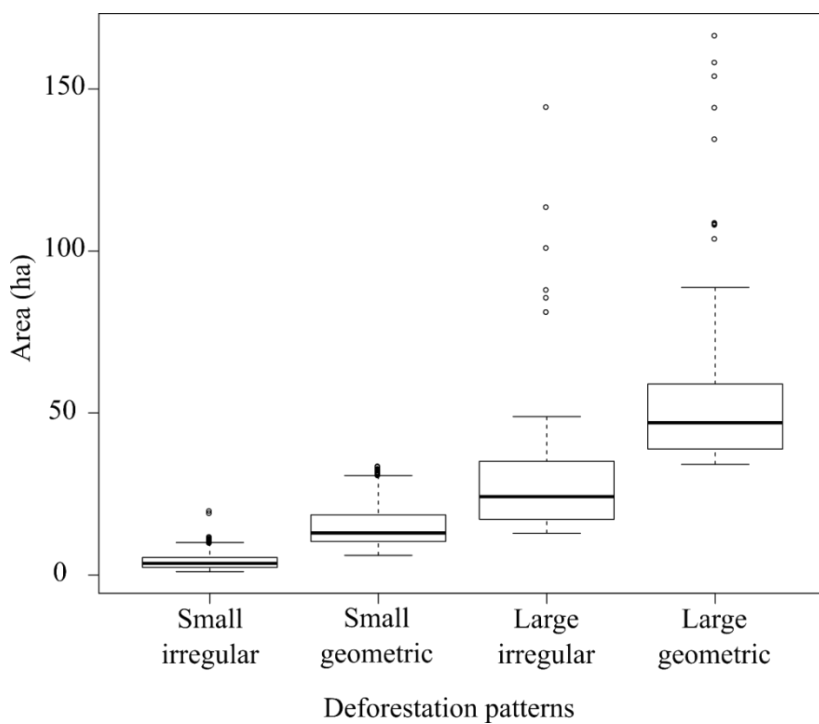


Fig. S4. Distribution of patch areas for each deforestation pattern.

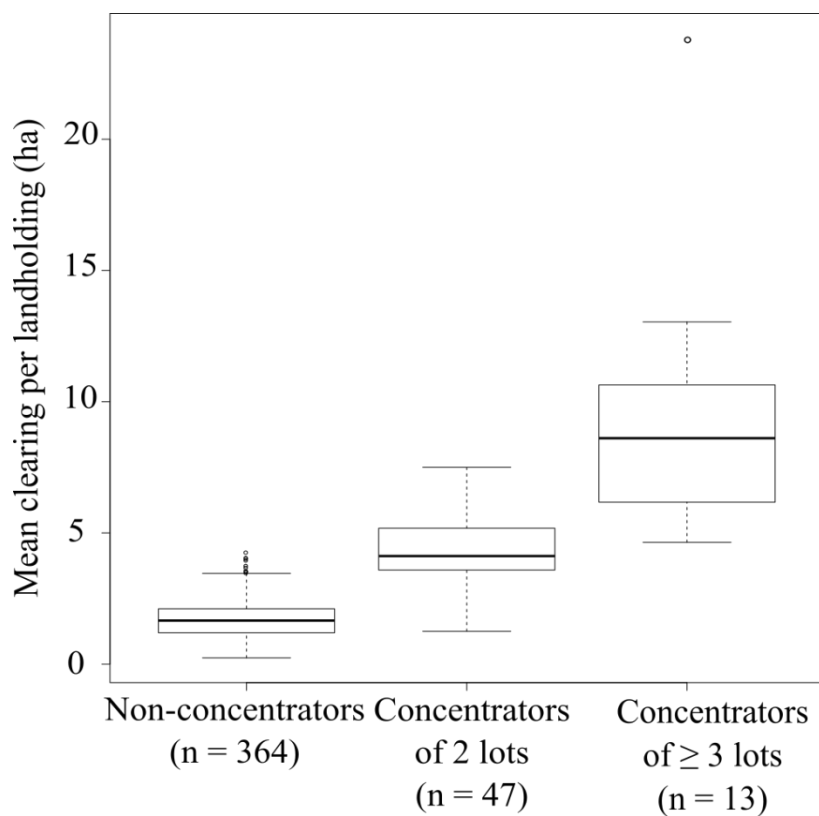


Fig. S5 Distribution of mean clearing per landholding from 1995 to 2011 separated into three groups: non-concentrators (n = 364 landholders), concentrators of 2 lots (n = 47 landholders) and concentrators of ≥ 3 lots (n = 13 landholders).

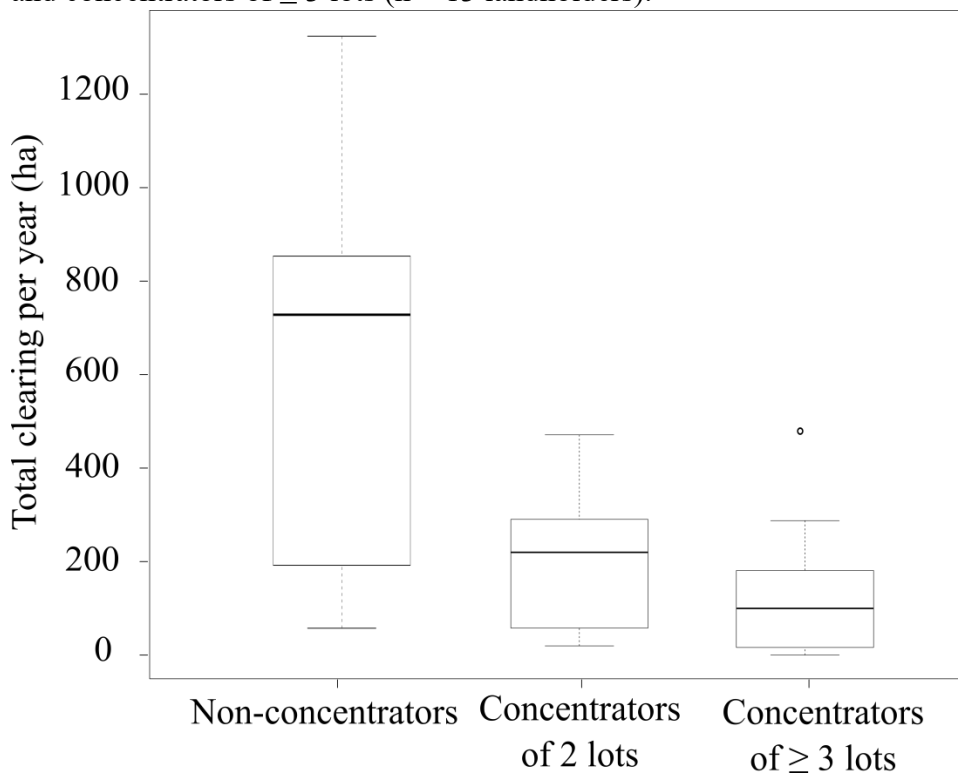


Fig. S6 Total area cleared in the period from 1995 to 2011 by type of actor (n = 17 years).

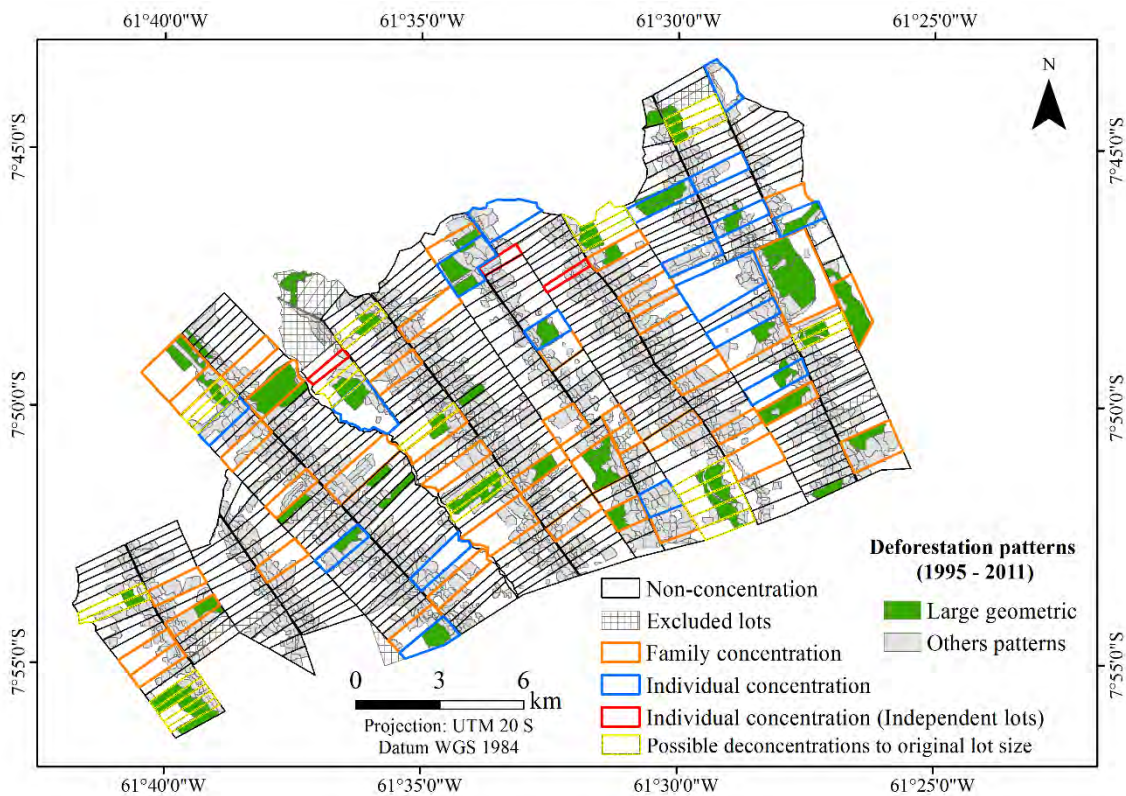


Fig. S7. The Matupi settlement indicating 30 lots held by non-concentrators that had portions of large geometric deforestation patches spanning more than one lot in 2011. This suggests fragmentation of previously concentrated landholdings.

References:

Brazil, INPE (Instituto Nacional de Pesquisas Espaciais), 2015. Nota Técnica – Correção da máscara do PRODES. http://www.obt.inpe.br/prodes/NT_deslocamentoMascara.pdf. (Last access 7 July 2017).

Brazil, INPE (Instituto Nacional de Pesquisas Espaciais), 2018. Projeto PRODES - Monitoramento da Floresta Amazônica Brasileira por Satélite. INPE, São José dos Campos, SP, Brazil. <http://www.dpi.inpe.br/prodesdigital/> (Last access 8 June 2018).

Brazil, INPE (Instituto Nacional de Pesquisas Espaciais), 2019. TerraBrasilis. Available at: <http://terrabrasilis.dpi.inpe.br/>. (Last access 31 January 2020).

Friedl, M.A., Brodley, C.E., 1997. Decision tree classification of land cover from remotely sensed data. *Remote Sens. Environ.* 61, 399–409. [https://doi.org/10.1016/S0034-4257\(97\)00049-7](https://doi.org/10.1016/S0034-4257(97)00049-7)

Körting, T. S., Fonseca, L. M. G., Câmara, G., 2013. GeoDMA—Geographic data mining analyst. *Comput. Geosci.* 57, 133-145. <https://doi.org/10.1016/j.cageo.2013.02.007>

Quinlan, J., 1993. C4.5: Programs for Machine Learning. Morgan Kaufmann, San Mateo, CA. Available at:

https://books.google.com.br/books?id=HExncpjbYroC&printsec=frontcover&hl=pt-BR&source=gbs_ge_summary_r&cad=0#v=onepage&q&f=false

Quinlan, J.R., 1996. Learning decision tree classifiers. *ACM Comput. Surv.* 28, 71–72. <https://doi.org/10.1145/234313.234346>

Walker, W.S., Gorelik, S.R., Baccini, A., Aragon-Osejo, J.L., Josse, C., Meyer, C., Macedo, M.N., Augusto, C., Rios, S., Katanh, T., de Souza, A.A., Cuellar, S., Llanos, A., Zager, I., Mirabal, G.D., Solvik, K.K., Farina, M.K., Moutinho, P., Schwartzman, S., 2020. The role of forest conversion, degradation, and disturbance in the carbon dynamics of Amazon indigenous territories and protected areas. *Proceedings of the National Academy of Sciences of the USA.* 117(6), 3015–3025. <https://doi.org/10.1073/pnas.1913321117>